

**APPLICATION OF SEPARABLE PROGRAMMING
TO REGIONAL WATER QUALITY MANAGEMENT**

by

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CHAPTER 1

INTRODUCTION

Often interrelated and competing uses of water give rise to complex water quality problems, especially in areas where the available water resources are hardly adequate enough to meet the growing demands. One such area is the Utah Lake drainage area, characterized by growing urbanization and also by extensive water uses for agricultural purposes. Water quality deterioration is caused by both urban and agricultural users. However, the nature of the water quality problems posed by urban and agricultural return flows, respectively, are quite distinct. Urban return flows contain large amounts of bacterial wastes and require treatment to reduce the biochemical oxygen demand (BOD). On the other hand, agricultural return flows contain large amounts of dissolved salts, and pose a salinity problem.

It is quite apparent that any efforts to maintain water quality in the area should aim at coordinating pollution abatement strategies among both types of users, the required water quality standards being met at minimum total costs. This involves setting up a mathematical model for the urban-agricultural water quality control system.

Huntzinger (1971) has carried out a comprehensive study on the future water usage models of Utah Valley to determine the shortages and surpluses of water under present conditions, and those resulting from changing future demands. A number of water management alternatives have been analyzed to determine their relative worth in efficiently

utilizing water supply to satisfy present and future water demands. The mathematical model used in this study was originally proposed by Walker, et al., (1973a, 1973b). A number of researchers, (Smith, 1968, Michel, 1970), have developed cost functions for the components of the system. Details of the system flows as well as the associated salt concentrations are given by Walker, et al. (1973b). Walker, et al., (1973a) have solved the resulting optimization problem using the 'Jacobian differential algorithm'. Shojalashkari (1974) has used the generalized reduced gradient method in optimizing the same system. In this study, another simpler nonlinear optimization technique, separable programming, has been used in optimizing the system. The resulting optimal policies are discussed and compared with earlier solutions of Walker, et al., and Shojalashkari. A review of the report by Fuhriman, Merritt, et al. (1975) is also given.

The contents of the report will be briefly presented in the following few paragraphs. The Utah Lake drainage area has been divided into four districts. A brief description of the area and its characteristics is given in the first chapter. The water demands in the area are also presented.

A mathematical model proposed originally by Walker, et al., for coordinating pollution control activities in the urban and agricultural sectors, is discussed. The cost functions associated with the components of the system are also presented. The objective function for the optimization problem and the constraints are then developed. Optimal policies, assuming independent operation of the districts, are determined

using separable programming. The results obtained are compared with those of Shojalashkari using the generalized reduced gradient method. A review of the report by Fuhriman, Merritt, et al. follows.

The next chapter explores the benefits resulting from coordinating pollution control strategies among the districts. The so called regional model is developed. The optimal policies are presented and discussed. Once again the results obtained are compared with earlier optimal policies obtained by Shojalashkari.

The following chapter deals with a discussion of the results presented by Walker, et al. (1974). The problem as presented by Walker, et al. was resolved using separable programming. The resulting optimal policies are presented and compared with those of Walker, et al.

Although the presentation of the model and its findings are carried out in the context of the Utah Lake drainage area, the approach is general enough to be applicable to other similar situations.

CHAPTER 2

THE UTAH LAKE DRAINAGE AREA

2.1. DESCRIPTION OF THE AREA

The Utah Lake drainage area is part of the Great Salt Lake drainage area, located in the north central part of Utah, as shown in Fig. 1. It encompasses about 3,356 square miles, and covers a major part of Utah County in addition to areas in four other counties, Sanpete, Juab, Wasatch, and Summit. The drainage area can be hydrologically divided into subareas, as shown in Fig. 2 (Huntzinger, 1971). This study concentrates on the Utah Valley.

As proposed by Walker, et al. (1973b) (see Fig. 3), the Utah Valley drainage area is divided into four districts, namely the Lehi-American Fork District, the Provo District, the Spanish Fork District, and the Elberta-Goshen District. They are being supplied by the American Fork River, Provo River, Spanish Fork River, and Currant Creek, respectively. It is noted that the return flows from Northern Juab Valley constitute the supply source for Currant Creek, and that the Elberta-Goshen District includes the western part of Goshen Valley, which is not supplied by the Spanish Fork River system.

The return flows from the four districts of the Utah Valley are eventually discharged into Utah Lake. Utah Lake is shallow (average depth about 8 ft., maximum depth about 20 ft.), has gently sloping shores, and is almost 19 miles long in the north-south direction, and about 10 miles wide (Huntzinger, 1971). To the east of the lake is

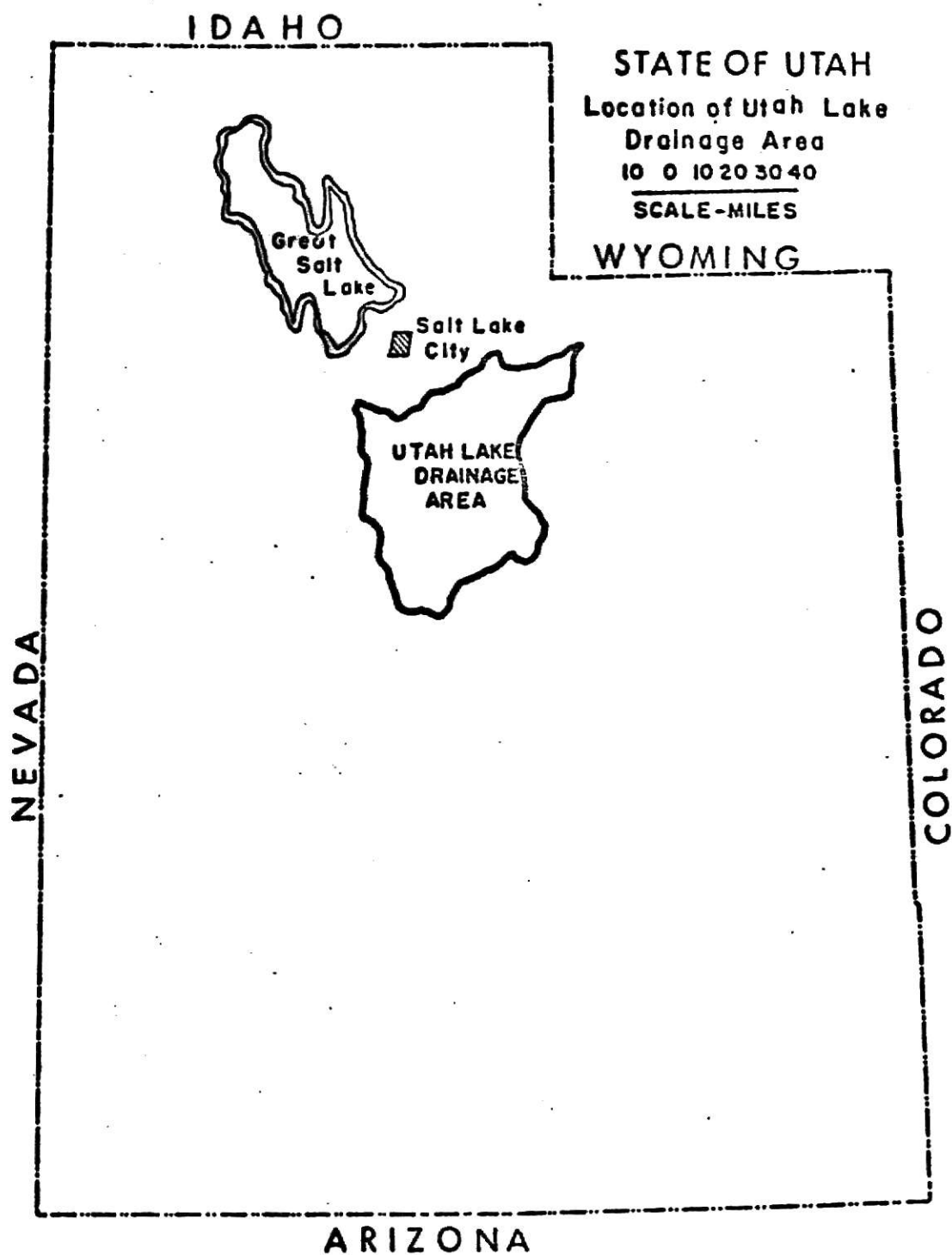


Fig. 1. Location of the Utah Lake drainage area [Huntzinger, 1971]

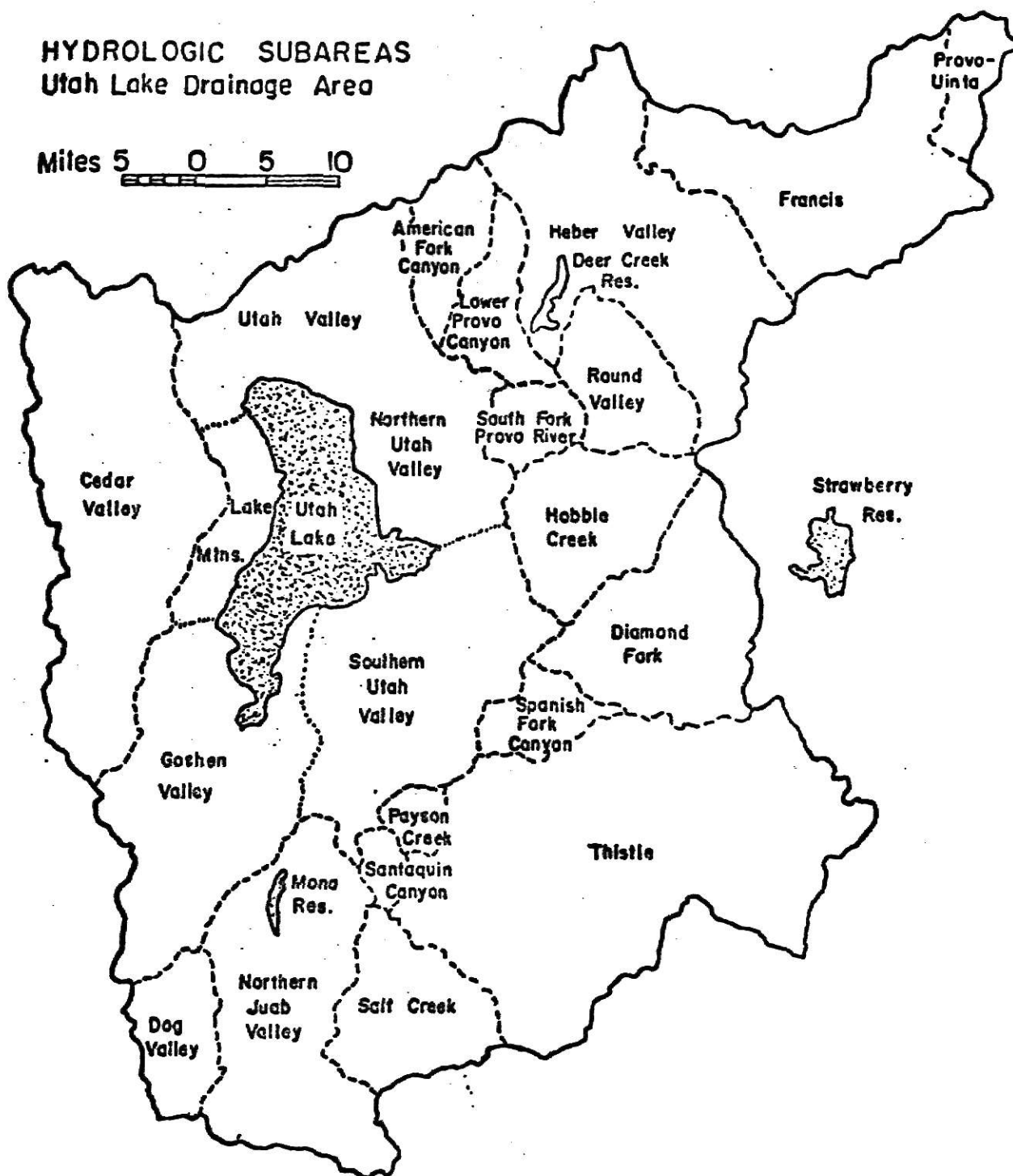


Fig. 2) Hydrologic subareas of the Utah Lake drainage area
(Huntzinger, 1971).

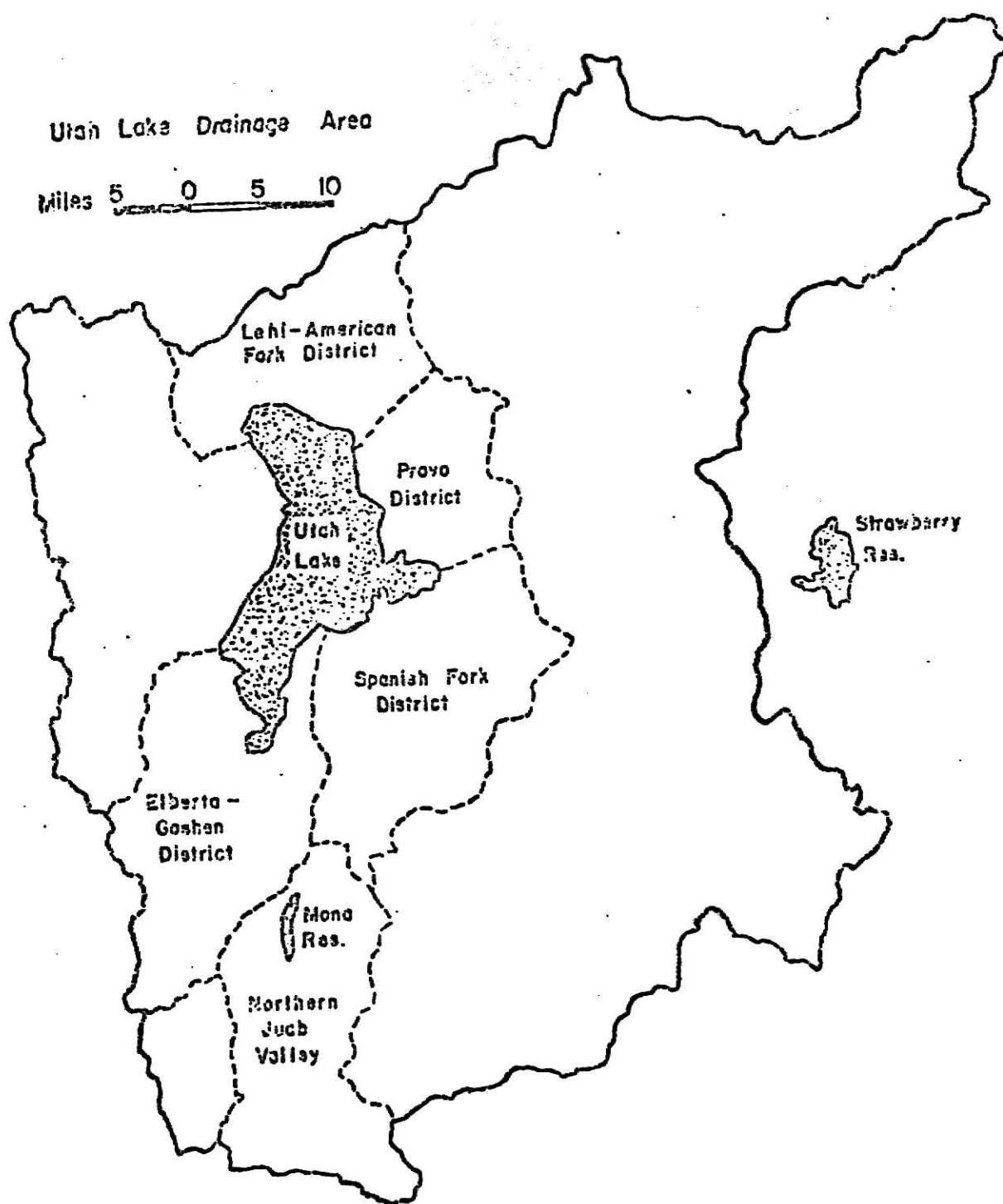


Fig. 3. General map of the Utah Lake drainage area (Huntzinger, 1971).

located a swampy area called Provo Bay which is connected to the Lake through a channel. Utah Lake drains into Great Salt Lake through its outlet, the Jordan River, which runs in a northerly direction.

The Utah Valley has a temperate and arid climate; low rainfall and humidity, high evaporation rate; mild summers and cold winters are also characteristics of the area (Huntzinger, 1971).

2.2. WATER RESOURCES AND DEMANDS IN THE UTAH VALLEY

The major portion of the available water resources in the Utah Lake drainage area are a number of streams and storage reservoirs. In Table 1 the major stream flows in the Utah Lake drainage area are listed. Two major streams of the area are the Provo River with a mean annual flow of 352,200 ac. ft., and the Spanish Fork River with a mean annual flow of 268,200 ac. ft.

The stream flow regulation in the area has been achieved through a number of small reservoirs, principal among them being: 1) the Deer Creek Reservoir at the lower end of the Heber Valley. This reservoir supplies the municipal and industrial water demands in Salt Lake County, and discharges about 96,700 ac. ft. per year into the Provo River (Hyatt, et al., 1969); 2) the Strawberry Reservoir in the Uinta Basin which is the source of the interbasin water transfer into the Diamond Fork River through the Strawberry Tunnel; and 3) the Mona Reservoir at the northern edge of the Northern Juab Valley. It holds the return flows of the Northern Juab Valley and Carrant Creek, and supplies the water needs of the Elberta-Goshen District. In addition to surface water resources, there are six major ground water sources in the Utah Lake drainage area

Table 1. Mean annual flow of major streams in the Utah Lake drainage area [Hyatt, et al., 1969].

River	Mean Annual Flow, acre-feet
Provo River	
near Kamas	34,300
Duchesne Tunnel	37,200*
Weber-Provo Diversion Canal	56,200*
at Hailstone	214,500
Ontario Tunnel	10,000*
Dry Creek and Fort Creek	20,000
American Fork River	38,200
Battle Creek	4,000
Grove Creek	3,000
Rock Creek	8,000
Hobble Creek	29,500
Spanish Fork River	
at Thistle	56,400
Strawberry Tunnel	60,800*
at Castilla	151,400
Payson Creek	9,400
Summit Creek	8,900
Salt Creek near Nephi	19,300
Currant Creek below Mona Reservoir	15,000
Jordan River	261,000

* Interbasin Transfer

(Walker, et al., 1973b). These are Kamas Valley, Heber Valley, Cedar Valley, Northern Juab Valley, Northern Utah Valley, and Southern Utah Valley, which includes Goshen Valley.

Agricultural activities in the Utah Valley are the single most important source of water use. Of a total of 219,658 acres of cultivated land in the area, about 74% is irrigated. The rest is dry farming. The estimated agricultural water demands under the 1970 conditions were 97,100 ac. ft. for the Lehi-American Fork District, 158,100 ac. ft. for the Provo District, 204,600 ac. ft. for the Spanish Fork District, and 35,600 ac. ft. for the Elberta-Goshen District. Table 2 lists the total agricultural water demands and supplies under the 1970 conditions (Walker, et al., 1973b).

Municipal and industrial water demands in the area, as of the present, constitute only about one-third of the agricultural demands. However, they are expected to increase in the future as urbanization and industrialization in the area increase, and agricultural activities decline. Table 3 shows the present and future urban water demands in the Utah Valley. The data are based on the projections made by the Economic Research Service of the U. S. Department of Agriculture (Huntzinger, 1971).

2.3. POLLUTION SOURCES IN THE UTAH VALLEY

As stated earlier, agricultural activities which are mostly concentrated in the Southern Utah Valley are the major sources of water use in the area. Urbanization in this area has been slow compared to the Northern Utah Valley where population concentration is high. This

Table 2. Annual agricultural water demands in the area under the 1970 conditions [Walker, et al., 1973b]
 Figures in ac. ft.

Districts	Supply Sources			Total Demand	Actual Agricultural Consumption	Agricultural Return Flows
	Canal Diversion	Precipitation	Pumped Water			
Lehi-American Fork	52,600	17,500	6,600 20,400*	97,100	39,800	57,300
Provo	131,100	19,000	8,000	158,100	74,300	83,800
Spanish Fork	135,400	51,400	17,800	204,600	100,300	104,300
Elberta-Goshen	13,500	10,100	1,200 10,800**	35,600	16,100	19,500

* Pumped water from the Provo District

**Pumped water from the Spanish Fork District

Table 3. Annual urban water demands in the Utah Valley area (Huntzinger, 1971), present and future. Figures in ac. ft.

District	1960		1980		2000		2020	
	Municipal	Industrial	Municipal	Industrial	Municipal	Industrial	Municipal	Industrial
Lehi-American Fork	6,500	1,000	10,700	1,500	17,000	2,200	23,000	3,000
Provo	18,000	64,000	29,800	106,000	48,000	130,000	65,600	144,000
Spanish Fork	2,000	6,000	5,000	7,500	6,700	10,000	7,800	15,000
Elberta-Coshen	250	---	350	---	500	---	750	---

gives rise to water quality problems that are mostly urban in nature in the northern part and agricultural in the southern part of the Valley. However, at the present time, the urban return flows pose a less serious pollution problem than the agricultural return flows with their high concentration of dissolved solids, which do represent a salinity control problem.

Utah Lake is the receiver of urban and agricultural return flows from all the districts. The Lake, in turn, drains into Great Salt Lake by passing through the Jordan River which is located in Utah County and which supplies most of the water needs of the Salt Lake City metropolitan area. Evaporation rate of the Lake being rather high, the salt concentration of the flows leaving the Lake are almost twice that of the inflows (Walker, et al., 1973b), posing a serious salinity problem in the Jordan River. The problem is only expected to become more acute in the future as water demands in the Salt Lake metropolitan area are expected to increase sharply. There is increasing possibility that additional water supplies from the Utah Lake drainage area will be utilized to meet growing demands. Any water quality control program must therefore necessarily address itself not only to the question of urban waste management, but also to the problem of salt concentration in the Utah Lake drainage area.

Water quality management in the Utah Valley can be achieved by any one of several possible approaches. One approach would be to impose quality standards on the Utah Lake or its outlets; another would be to set effluent standards on the return flows entering the Lake. Yet

another alternative approach to agricultural salinity control can be initiated by imposing standards on the sources of agricultural water.

In this regard, it must be pointed out that as a result of the anticipated growth of urbanization in the Utah Valley and therefore the need for reallocation of the existing water resources among the competing demands, the state of Utah has devised the Central Utah Project (Walker, et al., 1973a, 1973b). The main features of the project are water transfers from the Colorado River Basin into the Utah Valley, and from the Utah Valley itself to Salt Lake County, and provisions for reducing the evaporation rate in Utah Lake by separating the Goshen and Provo Bays from the main body of the Lake, thus reducing the surface area of the Lake.

CHAPTER 3

THE DISTRICT SALINITY - BOD CONTROL MODEL

This section deals with the development of a mathematical model for representing the salinity-BOD control system. A description of the various system components is given, followed by a discussion of the cost functions for the components. The objective function and the constraints imposed on the model are then defined. The optional policies obtained under the assumptions of individual operation of the districts are presented, followed by a comparison with earlier solutions obtained by Shojalashkari (1974). The report by Fuhriman, Merritt, et al., (1975) is reviewed.

For the purposes of this study, the water quality criteria to be controlled are the total dissolved solids (TDS) and the biochemical oxygen demand (BOD) concentrations. Although the waste treatment facilities remove such pollutants as suspended solids, phosphate, etc., the water quality standards here are limited to TDS and BOD as the two parameters that best characterize the quality (Walker and Skogerboe, 1973), and that are commonly measured and used in the design and monitoring of water quality schemes.

3.1. THE SYSTEM COMPONENTS

Figure 4 shows the schematic diagram of the urban-agricultural salinity-BOD control system for a district (Walker, et al., 1973b). In any district an urban sector and an agricultural sector are distinguished. The urban sector waste water flows, Q_u in million gallons per day (mgd),

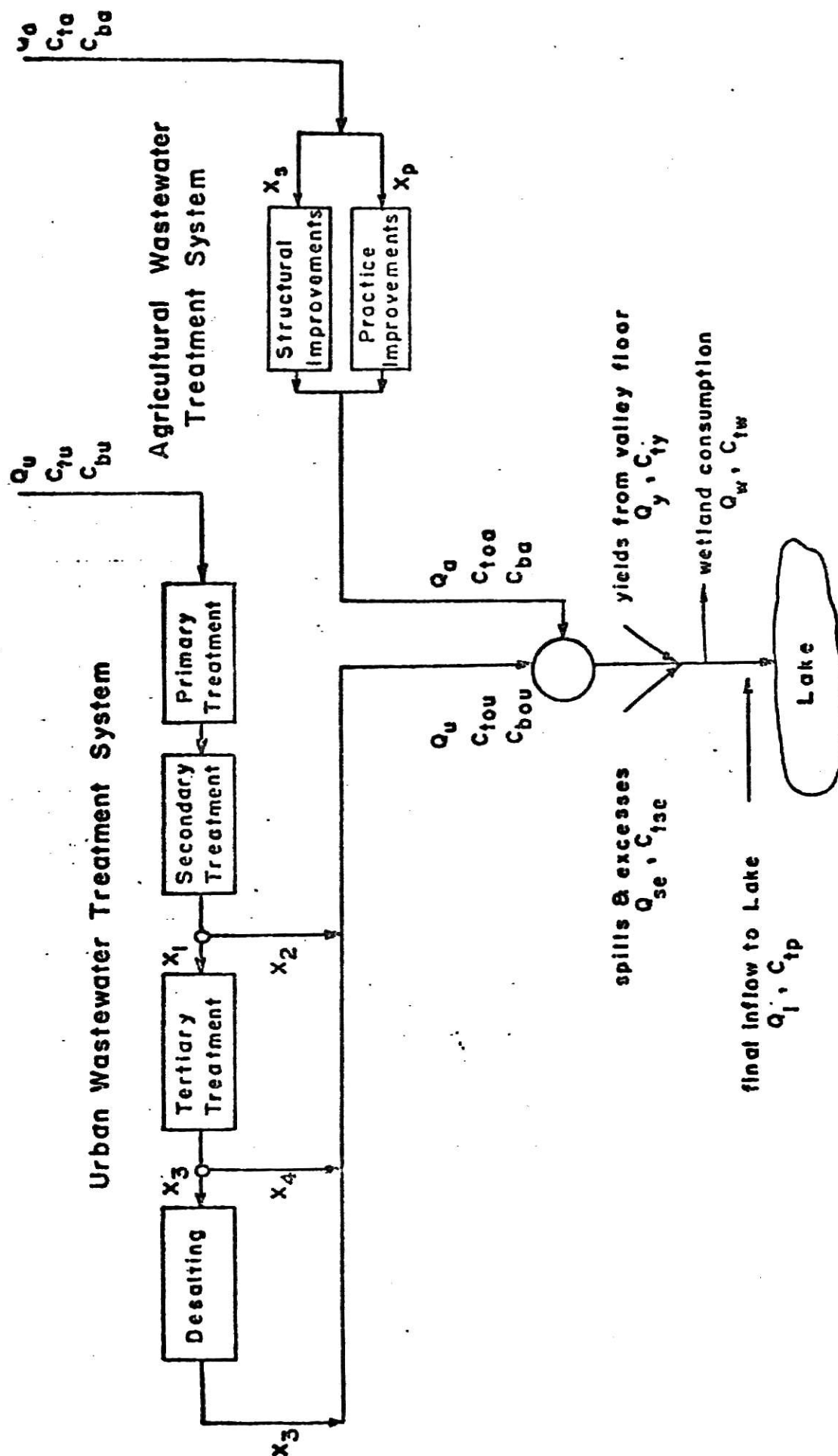


Fig. 4. Schematic diagram of the return flows in a district [Walker, et al., 1973b].

are collected and directed to the primary-secondary treatment unit. These flows have a TDS content of C_{tu} mg/l and a BOD content of C_{bu} mg/l.

At the outlet of the primary-secondary treatment unit, the outflows are divided into two streams, of which X_1 mgd are sent into tertiary treatment for further BOD removal, and X_2 mgd are directed to the effluent channel of the sector. At the outlet of the tertiary unit, of X_1 mgd, X_3 mgd are fed into the desalting plant, and X_4 mgd are directed to the effluent channel. As a result of the various treatments, the effluents from the urban sector have a lower TDS concentration of C_{tou} and a lower BOD content of C_{bou} .

In the agricultural sector it is assumed that BOD concentration of the irrigation flows, C_{ba} , are low enough so that no BOD reduction would be required. On the other hand, the salt content of the agricultural return flows is normally high. Hence salt reduction is required. Salt reduction in the agricultural sector does not involve physical removal of salts in a desalting plant. This would be very expensive as the water has to undergo tertiary treatment before it can be treated in the desalting plant. Otherwise, it fouls up operation of the plant. Irrigation flows are prevented from picking up salt through two kinds of practices (Walker and Skogerboe, 1973):

- 1) Structural improvements, that is, physical improvements of the irrigation system which reduce the amount of salt picked up by the irrigation flows, employing such measures as canal lining, land leveling, measuring and control structures, delivery of water to the fields

only when needed, and in amounts to the demands, even distribution of water into the soil, etc.

ii) Practice improvements, that is, correcting the effects of poor water management. Included under this category are such practices as minimization of surface areas by eliminating unnecessary ponds and marshes, reduction of salt leaching from the soil and granular aquifers by adopting different methods of irrigation in different areas, and reducing the seepage and dumping of surplus flows into natural wasteways through tighter control of water in the conveyance system.

As a result of the two salinity-control practices described above, the salt content of the agricultural return flows is reduced to C_{toa} mg/l.

The urban and agricultural return flows are then directed into Utah Lake along with three other flows. Spills and excess flows, Q_{se} , in a district are those flows that are not diverted to urban and agricultural uses. Associated with the spills and excesses is a salt content of C_{tse} mg/l. Yields from the valley floor, Q_y , are those flows that originate as surface runoff and deep percolation from precipitation on non-agricultural lands. A TDS concentration of C_{ty} mg/l is associated with these flows. Finally, the wetland consumptive uses are the potential wetland demands less the precipitations in the concerned district. The sum of the agricultural and urban return flows, yields from the valley floor, and spills and excesses minus the wetland consumptive use constitute the total inflow to the Lake, Q_I , from a district.

Figure 5 represents the water and salt balances in the Utah Valley under the 1970 conditions (Walker, et al., 1973b). Table 4 summarizes these data. The data have been compiled and in some cases estimated by Walker and his associates, based on a number of sources including the U. S. Bureau of Reclamation (1964a, 1964b, 1964c), Hyatt, et al., (1969), and Huntzinger (1971). The reader is referred to Walker, et al., (1973b) for details on the methods of analysis and estimation of data as well as the assumptions made to construct the water and salt budget diagram of Fig. 5.

3.2. COST FUNCTIONS FOR THE SYSTEM COMPONENTS

The objective of the district model is to achieve a desired level of salinity-BOD concentration in the return flows from the district at a minimum total annual cost. This section deals with the cost functions used in setting up the objective function.

Beginning with the urban sector, the waste treatment cost information is reported by several researchers (for example, Smith, 1968; Michel, 1970; Shah and Reid, 1970). The cost figures represent the capital costs of equipment as well as operating and maintenance costs.

In what follows, Y_c represents the total capital costs for the equipment in question, in terms of millions of dollars, Y_{om} the operating and maintenance costs in terms of ¢/1000 gallons of water treated, and Z the plant capacity in million of gallons per day (mgd).

For the primary treatment unit, Smith (1968) suggests the following cost functions in terms of 1967 dollars:

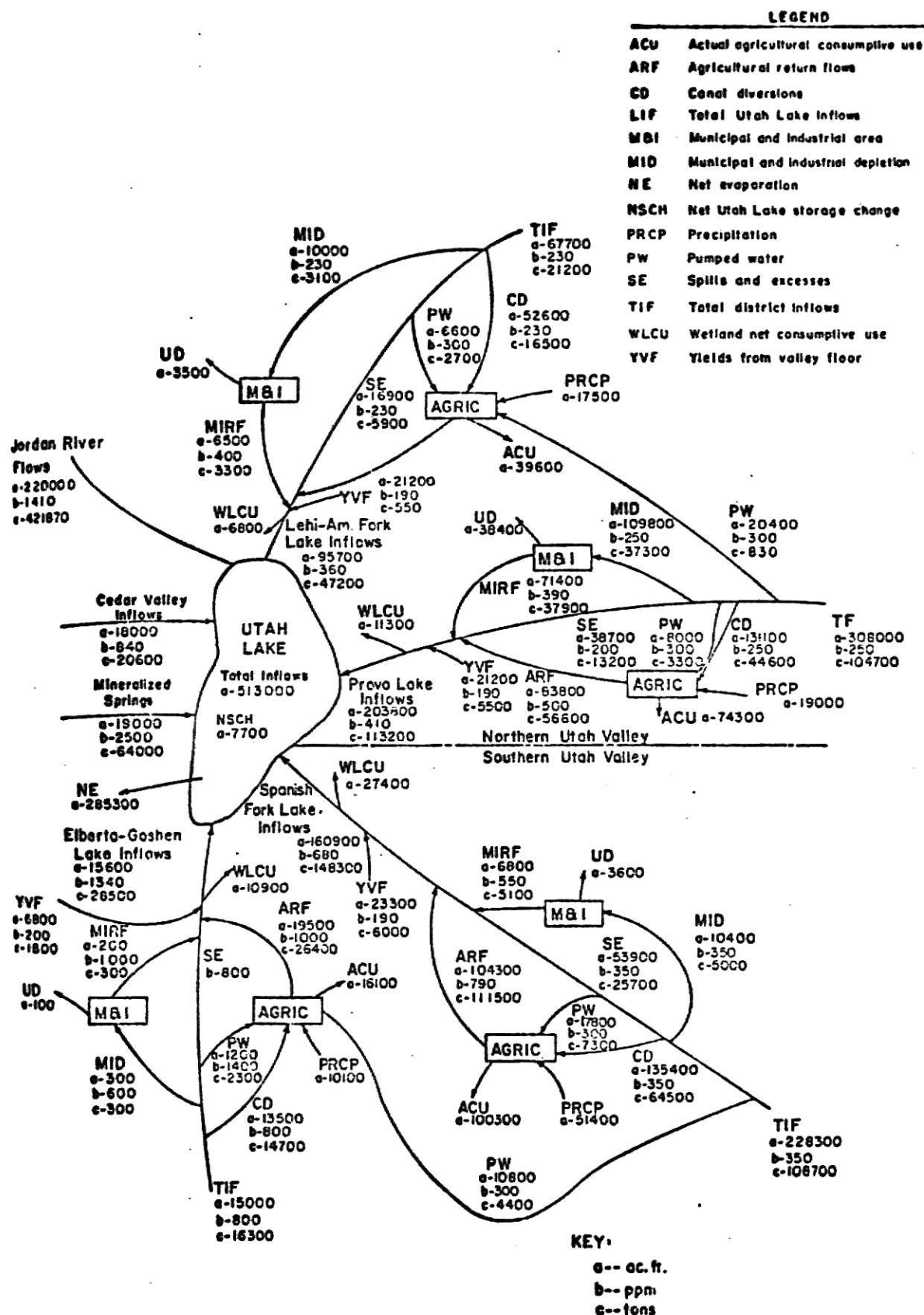


Fig. 5. Water and salt balance diagram for Utah Valley under 1970 conditions [Walker, et al., 1973b].

Table 4. Water and Salt Budget for Utah Valley under 1970 Conditions

Districts	Flows*						Salt Contents, mg/L					BOD Contents, mg/L	
	Urban Sector Q_u	Agricultural Sector Q_a	Yields Q_y	Spills & Excesses Q_{se}	Wetland Consumption Q_w	Total Inflow Q_I	Urban C_{tu}	Agric. C_{ta}	Yields C_{ty}	Spills & Excesses C_{tse}	Present Values C_{tp}	Urban C_{bu}	Agric C_{ba}
Lehi-American Fork	6,500.0 6.42	57,300.0 56.58	21,200.0 20.93	17,500.0 17.28	6,800.0 6.71	95,700.0 94.50	400	420	190	230	362	180	15
Provo	71,400.0 70.51	83,600.0 82.75	21,200.0 20.93	38,700.0 38.22	11,300.0 11.16	203,800.0 201.25	390	500	190	200	400	180	15
Spanish Fork	6,800.0 6.71	104,300.0 103.0	23,300.0 23.01	53,900.0 53.23	27,400.0 27.06	160,900.0 158.89	550	790	190	350	680	180	15
Elberta-Goshen	200.0 0.20	19,500.0 19.26	6800.0 6.71	0.0 0.0	10900.0 10.76	15,600.0 15.40	1000	1000	200	0.0	1350	185	15

* Figures in the first row are flows in terms of ac. ft., and in the second row in terms of mgd.

** Assumed figures based on information provided by the Utah State Health Division (1974).

$$Y_c = 0.316 Z^{0.71} \quad (1)$$

$$Y_{om} = 4.47 Z^{-0.17} \quad (2)$$

For the secondary treatment unit, the following capital cost function in terms of 1967 dollars, is given by Smith (1968):

$$Y_c = 0.58 Z^{0.80} \quad (3)$$

For the same unit the operating and maintenance cost function, again in 1967 dollars, is given by Michel (1970):

$$Y_{om} = 9.02 Z^{-0.107} \quad (4)$$

The tertiary treatment unit consists of three components each: flocculation, lime treatment, and sedimentation; granular carbon adsorption; and finally ammonia stripping. Each component has capital costs, and operating and maintenance cost functions. Smith (1968) has suggested the following cost functions in terms of 1967 dollars:

$$Y_c = 0.05 Z^{0.89} \quad (5)$$

$$Y_{om} = 2.99 Z^{-0.038} \quad (6)$$

for flocculation, lime treatment, and sedimentation;

$$Y_c = 0.398 Z^{0.65} \quad (7)$$

$$Y_{om} = 10 Z^{-0.28} \quad (8)$$

for granular adsorption; and

$$Y_c = 0.0398 Z^{0.90} \quad (9)$$

$$11.58 Z^{-0.30} \quad Z \leq 3 \text{ mgd} \quad (10)$$

$$Y_{om} = 1.20 Z^{-0.04} \quad Z > 3 \text{ mgd} \quad (11)$$

for ammonia stripping.

The desalting plant is assumed to be of an electrodialysis type for which the following cost functions, in 1967 dollars, are suggested by Smith (1968):

$$Y_c = 0.51 Z^{0.67} \quad (12)$$

$$Y_{om} = 47.94 Z^{-0.21} \quad (13)$$

which are for a 4-stage demineralization system with a removal efficiency of 90% (Walker and Skogerboe, 1973).

In the agricultural sector, the costs are of a different nature. As mentioned earlier, the salt pick-up by the irrigation flows is minimized through structural improvements and practice improvements. Walker and Skogerboe (1973) have estimated the relationship between the percentage of salt that can be reduced through the above-mentioned practices, and the percentage of area treated. These relationships are represented in Fig. 6. For example, if 50% of the entire district were to be rehabilitated by both structural improvements and improved irrigation practices, 30% of the salt can be removed through structural improvements and 44% through practice improvements, for a total TDS removal of 74%. If the entire area were to be rehabilitated by both

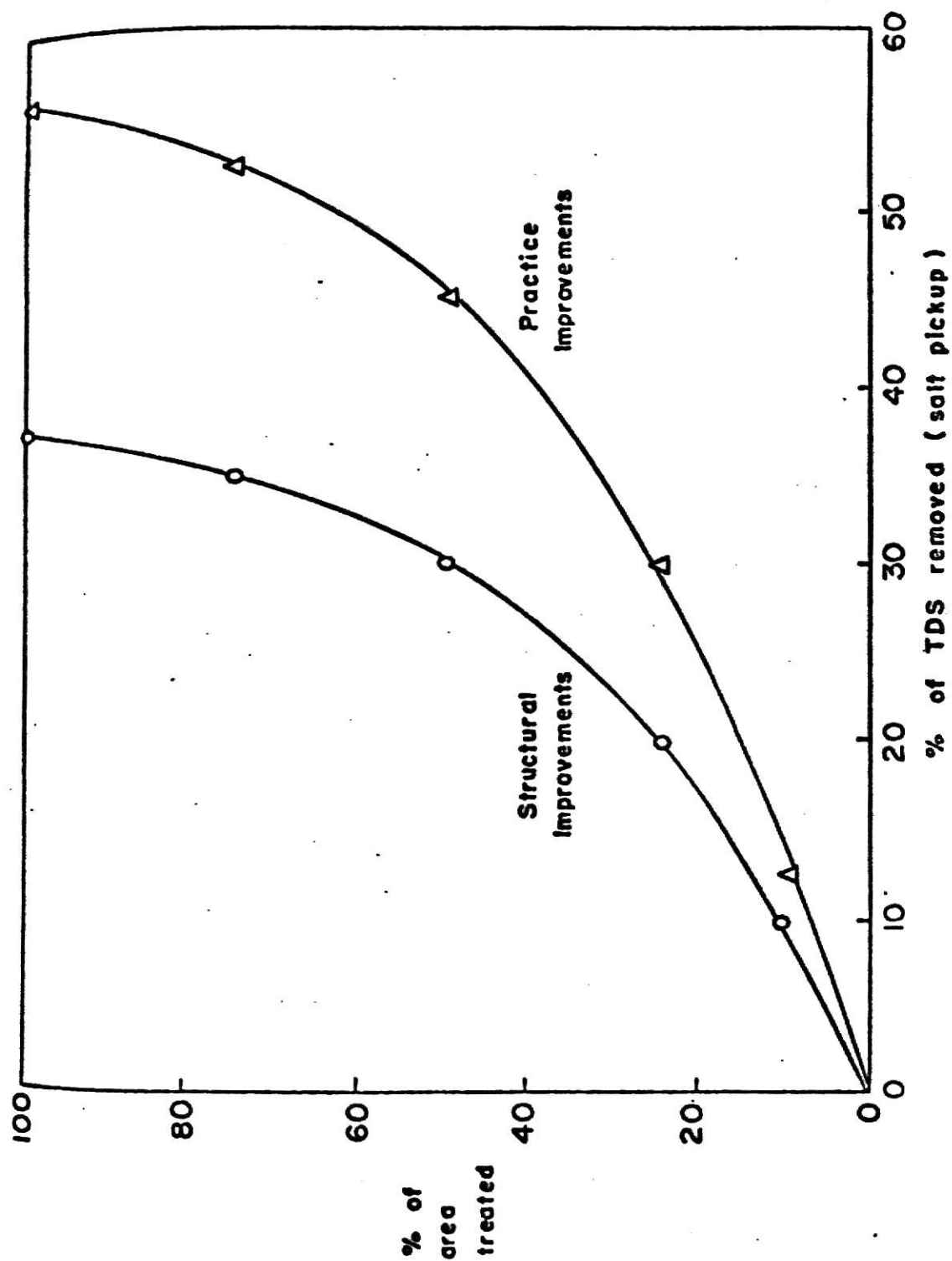


Fig. 6. Cost-effectiveness distributions for agricultural salinity control [Walker, et al., 1973b].

methods, the percentages of salt removed are 37.5% and 56.5% through structural and practice improvements, respectively, and total TDS removed is 94%. Walker, et al., (1973b) have estimated the total annual costs when the entire area in each district is to be rehabilitated through each of the two practices. The salt removal cost through each practice is proportional to the percentage of the area treated.

Let C_s and C_p be the total annual cost of rehabilitating the entire area of a district through structural and practice improvements, respectively. Also, let X_s represent the percentage of the TDS removed through structural improvements. This requires that $f_s(X_s)$ percent of the area be rehabilitated by this method. Similarly, X_p percent removal of TDS by improved practices requires that $f_p(X_p)$ percent of the area be treated by this method. The relationships between $f_s(X_s)$ and X_s , and between $f_p(X_p)$ and X_p are given in Fig. 6. If the percentages of salt reduction through structural and practice improvements are chosen as X_s^* and X_p^* respectively, the percentages of the area treated through each method would be $f_s(X_s^*)$ and $f_p(X_p^*)$, respectively, and the total annual costs of agricultural salinity control in that district is

$$Y_A = C_s f_s(X_s^*)/100 + C_p f_p(X_p^*)/100 . \quad (14)$$

The functional forms of the relations $f_s(X_s)$ and $f_p(X_p)$ are estimated as follows:

$$f_s(X_s) = 34.42 [3.68135 - \ln(39.70 - X_s)], \quad X_s \leq 37.50 \quad (15)$$

$$f_p(X_p) = 37.60 [4.11087 - \ln(56.50 - X_p)], \quad X_p \leq 56.0 \quad (16)$$

The total annual costs of agricultural rehabilitation, C_s and C_p , are given in the Walker, et al. Report (1973b) and summarized in Table 5.

3.3. FORMULATION OF THE OBJECTIVE FUNCTION

This section and the following one deal with the formulations of the objective function and constraints respectively.

It is noted that in the preceding section the capital costs of the waste treatment facilities were expressed in terms of million dollars, the operating and maintenance costs in terms of ¢/1000 gallons of water treated. In both cases the costs were in 1967 dollars.

In order to convert all the costs into annual costs expressed in 1974 dollars it was assumed that the waste treatment facilities had a useful life of 30 years, that the interest rate for the purpose of capital recovery was 6% per year, and that a load-factor of 330 on-stream days per year would suffice. The costs were then increased by 44% to account for the rate of inflation since 1967 (Business Week, 1974). Under these assumptions the cost functions for the waste treatment equipment are:

1) Primary Treatment Unit

$$\begin{aligned} Y_c &= 1.44 \times (0.316 Z^{0.71}) \times 10^6 \times 0.07265 \\ &= 3.31 \times 10^4 Z^{0.71} \end{aligned} \quad (17)$$

where the capital recovery rate factor for a life of 30 years and an interest rate of 6% per year is .07265;

Table 5. Total Annual Agricultural Salinity Control Costs

District	C_s	C_p
Lehi-American Fork	\$704,840	\$2,544,270
Provo	914,510	2,984,570
Spanish Fork	1,302,060	7,855,670
Elberta-Goshen	527,260	1,698,070

$$Y_{om} = 1.44 \times (4.47 Z^{-0.17}) \left(\frac{1}{100} \frac{\$}{c} \right) \left(Z \frac{\text{million gallons}}{\text{day}} \right)$$

$$(10^6 \frac{\text{gallons}}{\text{mgd}}) \left(\frac{1}{1000} \times \frac{1}{\text{gallons}} \right) \times 330 \frac{\text{days}}{\text{year}}$$

$$= 1.44 \times 330 \times 10 \times (4.47 Z^{-0.17 + 1})$$

$$= 2.12 \times 10^4 Z^{0.83} \quad (18)$$

ii) Secondary Treatment Unit

$$Y_c = 1.44 \times 0.58 Z^{0.80} \times 10^6 \times 0.07265$$

$$= 6.07 \times 10^4 Z^{0.80} \quad (19)$$

$$Y_{om} = 1.44 \times 9.02 Z^{-0.107} \times 330 \times 10 Z$$

$$= 4.29 \times 10^4 Z^{0.893} \quad (20)$$

iii) Tertiary Treatment Unit

$$Y_c = 1.44 \times 10^6 \times (0.07265 \times 0.05 Z^{0.89} + 0.398 Z^{0.65} + 0.0398 Z^{0.90})$$

$$= 5.23 \times 10^3 Z^{0.89} + 4.16 \times 10^4 Z^{0.65} + 4.16 \times 10^3 Z^{0.90} \quad (21)$$

$$Y_{om} = (1.44 \times 330 \times 10 Z) \left\{ 2.99 Z^{-0.038} + 10 Z^{-0.28} \right. \\ \left. + \left\{ \begin{array}{ll} 11.58 Z^{-0.30} & Z \leq 3 \text{ mgd} \\ 1.20 Z^{-0.04} & Z > 3 \text{ mgd} \end{array} \right\} \right\}$$

$$\begin{aligned}
&= 1.42 \times 10^4 Z^{0.962} + 4.75 \times 10^4 Z^{0.72} \\
&+ \left\{ \begin{array}{ll} 5.50 \times 10^4 Z^{0.70}, & Z \leq 3 \text{ mgd} \\ 5.70 \times 10^3 Z^{0.96}, & Z > 3 \text{ mgd} \end{array} \right\} \quad (22)
\end{aligned}$$

iv) Desalting Unit

$$\begin{aligned}
Y_c &= 1.44 \times 0.51 Z^{0.67} \times 10^6 \times .07265 \\
&= 5.34 \times 10^4 Z^{0.67} \quad (23)
\end{aligned}$$

$$\begin{aligned}
Y_{om} &= 1.44 \times 47.94 Z^{-0.21} \times 330 \times 10 Z \\
&= 22.76 \times 10^4 Z^{0.79} \quad (24)
\end{aligned}$$

Referring to Fig. 4, the objective function for any district i can now be written as:

$$\begin{aligned}
Y_{D,i} &= 3.31 \times 10^4 (X_{1,i} + X_{2,i})^{0.71} + 2.12 \times 10^4 (X_{1,i} + X_{2,i})^{0.83} \\
&+ 6.07 \times 10^4 (X_{1,i} + X_{2,i})^{0.80} + 4.29 \times 10^4 (X_{1,i} + X_{2,i})^{0.893} \\
&+ 5.23 \times 10^3 X_{1,i}^{0.89} + 4.16 \times 10^4 X_{1,i}^{0.65} + 4.16 \times 10^3 X_{1,i}^{0.90} \\
&+ 1.42 \times 10^4 X_{1,i}^{0.962} + 4.75 \times 10^4 X_{1,i}^{0.72} \\
&+ \left\{ \begin{array}{ll} 5.50 \times 10^4 X_{1,i}^{0.70}, & X_{1,i} \leq 3 \text{ mgd} \\ 5.70 \times 10^3 X_{1,i}^{0.96}, & X_{1,i} > 3 \text{ mgd} \end{array} \right.
\end{aligned}$$

$$\begin{aligned}
& + 5.34 \times 10^4 X_{3,i}^{0.67} + 22.76 \times 10^4 X_{3,i}^{0.79} \\
& + C_{s,i} \times 34.42 \left[3.68135 - n (39.70 - X_{s,i}) \right] / 100 \\
& + C_{p,i} \times 37.60 \left[4.11087 - n (56.50 - X_{p,i}) \right] / 100 \\
& i = 1, 2, 3, 4.
\end{aligned} \tag{25}$$

3.4. CONSTRAINTS IMPOSED ON THE SYSTEM

The constraints on the system may be classified into two types: constraints on the physical flows in the system, and constraints on the BOD and TDS concentrations of the district effluents. Referring to Fig. 4, the following constraints on the physical flows are evident:

$$g_{1,i}(\bar{X}) = X_{1,i} + X_{2,i} - Q_{u,i} = 0 \tag{26}$$

$$g_{2,i}(\bar{X}) = X_{2,i} + X_{3,i} + X_{4,i} - Q_{u,i} = 0 \tag{27}$$

$$i = 1, 2, 3, 4.$$

In the formulation of the BOD and TDS constraints, the salt content C_{tou} and the BOD content C_{bou} can be calculated as follows:

$$\begin{aligned}
C_{\text{tou},i} &= \frac{(X_{2,i} + X_{4,i})C_{\text{tu},i} + (1-e_d) C_{\text{tu},i} X_{3,i}}{X_{2,i} + X_{3,i} + X_{4,i}} \\
&= \frac{(X_{2,i} + X_{4,i})C_{\text{tu},i} + (1-e_d) C_{\text{tu},i} X_{3,i}}{Q_{u,i}}
\end{aligned} \tag{28}$$

$$\begin{aligned}
 C_{bou,i} &= \frac{X_{2,i} (1-e_{ps})C_{bu,i} + (X_{3,i} + X_{4,i})(1-e_t)(1-e_{ps})C_{bu,i}}{X_{2,i} + X_{3,i} + X_{4,i}} \\
 &= \frac{X_{2,i} (1-e_{ps})C_{bu,i} + (X_{3,i} + X_{4,i})(1-e_t)(1-e_{ps})C_{bu,i}}{Q_{u,i}} \quad (29)
 \end{aligned}$$

where

e_d = salt removal efficiency of the desalting unit = 0.90

e_{ps} = BOD removal efficiency of the primary-secondary units
combined = 0.85

e_t = BOD removal efficiency of the tertiary unit = 0.99.

In the agricultural sector, the salt content of the return flows C_{toa} , is related to the percentages of salt removal by structural and practice improvements, $X_{s,i}$ and $X_{p,i}$, through the following relationship (Walker and Skogerboe, 1973):

$$X_{s,i} + X_{p,i} = 100 \times \frac{C_{ta,i} - C_{toa,i}}{C_{ta,i} - C_{tmin,i}} \quad (30)$$

where $C_{ta,i}$ is the salt content of the irrigation flows before the rehabilitation, and $C_{tmin,i}$ is the minimum salt content that can be achieved in district i. It may be recalled that the maximum possible salt reduction through structural and practice improvements is 94% of the salt picked up, $C_{tmin,i}$ would, therefore, be equal to 6% of $C_{ta,i}$. Hence, the value of $C_{toa,i}$ from eq. (30) is:

$$C_{toa,i} = C_{ta,i} - \frac{.94}{100} C_{ta,i} (X_{s,i} + X_{p,i}) \quad (31)$$

As no BOD treatment is assumed in the agricultural sector in this model, the original BOD concentration $C_{ba,i}$ remains unchanged.

The balance equations can now be written for the BOD and TDS concentrations at the point of entry into the Lake, to represent the BOD and TDS constraints:

$$Q_{u,i} C_{bou,i} + Q_{a,i} C_{ba,i} \leq (Q_{u,i} + Q_{a,i}) \times BOD_{std}, \quad i = 1, 2, 3, 4 \quad (32)$$

$$Q_{u,i} C_{tou,i} + Q_{a,i} C_{toa,i} + Q_{se,i} C_{tse,i} + Q_{y,i} C_{ty,i} \leq Q_{I,i} \times (C_{tp,i} - TDS_{std}), \quad i = 1, 2, 3, 4 \quad (33)$$

where BOD_{std} is the BOD standard imposed on the system outflows, $C_{tp,i}$ is the present concentration of salt in the district's effluents, TDS_{std} is the salt reduction standard imposed on the district's effluents, and $Q_{I,i}$ is the total inflow into the Lake from district i (see Table 4) and is defined as:

$$Q_{I,i} = Q_{u,i} + Q_{a,i} + Q_{se,i} + Q_{y,i} - Q_{w,i} \quad (34)$$

Substituting eqs. (28), (29), and (31) into eqs. (32) and (33) gives:

$$g_{3,i}(\bar{X}) = X_{2,i} (1-e_{ps}) C_{bu,i} + (X_{3,i} + X_{4,i}) (1-e_t) (1-e_{ps}) C_{bu,i} + Q_{a,i} C_{ba,i} - (Q_{u,i} + Q_{a,i}) \times BOD_{std} \leq 0 \quad (35)$$

$$\begin{aligned}
g_{4,i}(\bar{X}) = & (X_{2,i} + X_{4,i}) C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \\
& + Q_{a,i} C_{ta,i} - \frac{.94}{100} C_{ta,i} (X_{s,i} + X_{p,i}) \\
& + Q_{se,i} C_{tse,i} + Q_{y,i} C_{ty,i} - Q_{I,i} \times (C_{tp,i} - TDS_{std}) \leq 0 \\
& i = 1, 2, 3, 4 \quad (36)
\end{aligned}$$

The optimization problem for a district, therefore, is to minimize the total annual costs $Y_{D,i}$, as given by eq. (25), subject to the constraints, $g_{1,i}(\bar{X})$ through $g_{4,i}(\bar{X})$ as represented by eqs. (26), (27), (35), and (36).

3.5. APPLICATION OF SEPARABLE PROGRAMMING TO THE DISTRICT MODEL

Separable programming is a special class of nonlinear programming that can be adapted to linear programming. Separable programming problems must be of the form:

$$\begin{aligned}
& \text{maximize or minimize} \\
C(\bar{X}) = & \sum_{i=1}^m f_i(X_i) \quad (37)
\end{aligned}$$

subject to

$$\sum_{i=1}^m g_{ki}(X_i) \leq b_k, \quad k = 1, 2, \dots, P \quad (38)$$

and

$$X_i \geq 0, \quad i = 1, 2, \dots, m \quad (39)$$

It is noted that the objective function and the constraints are constructed of separable functions. Approximation of the nonlinear separable functions by piecewise linear functions yields a restricted linear programming problem which can be solved by a revised simplex method. MPS/360 has the necessary revision. For more details the reader is referred to Hadley (1964) and Williams (1972).

Often the nonlinear programming problem has some terms that are not separable functions. However, simple mathematical manipulations like taking logarithms or substitutions, yield separable functions. Once again the reader is referred to Hadley (1964) for more details.

It may be pointed out that in the optimization problem under discussion, the objective function, $Y_{D,i}$, represented by eq. (25) is not entirely made up of separable functions. However, it may be easily converted to the required form if we substitute

$$X_{7,i} = X_{1,i} + X_{2,i} \quad (40)$$

The separable programming problem is then:

Minimize

$$\begin{aligned} Y_{D,i} = & 3.31 \times 10^4 X_{7,i}^{0.71} + 2.12 \times 10^4 X_{7,i}^{0.83} + 6.07 \times 10^4 X_{7,i}^{0.80} \\ & + 4.29 \times 10^4 X_{7,i}^{0.893} + 5.23 \times 10^3 X_{1,i}^{0.89} + 4.16 \times 10^4 X_{1,i}^{0.65} \\ & + 4.16 \times 10^3 X_{1,i}^{0.90} + 1.42 \times 10^4 X_{1,i}^{0.962} + 4.75 \times 10^4 X_{1,i}^{0.72} \\ & + \begin{cases} 5.50 \times 10^4 X_{1,i}^{0.70} & , \text{ if } X_{1,i} \leq 3 \text{ mgd} \\ 5.70 \times 10^3 X_{1,i}^{0.96} & , \text{ if } X_{1,i} > 3 \text{ mgd} \end{cases} \end{aligned}$$

$$\begin{aligned}
& + 5.34 \times 10^4 X_{3,i}^{0.67} + 22.76 \times 10^4 X_{3,i}^{0.79} \\
& + C_{s,i} \times 34.42 \left[3.68135 - \ln (39.70 - X_{s,i}) \right] / 100 \\
& + C_{p,i} \times 37.60 \left[4.11087 - \ln (56.50 - X_{p,i}) \right] / 100 \\
& i = 1, 2, 3, 4
\end{aligned} \tag{41}$$

subject to the constraints

$$X_{1,i} + X_{2,i} = Q_{u,i} \tag{42}$$

$$X_{2,i} + X_{3,i} + X_{4,i} = Q_{u,i} \tag{43}$$

$$\begin{aligned}
& X_{2,i} (1-e_{ps}) X_{bu,i} + X_{3,i} (1-e_t)(1-e_{ps}) X_{bu,i} \\
& + X_{4,i} (1-e_t)(1-e_{ps}) X_{bu,i} \leq (Q_{u,i} + Q_{a,i}) \times BOD_{std} \\
& - Q_{a,i} C_{ba,i}
\end{aligned} \tag{44}$$

$$\begin{aligned}
& X_{2,i} C_{tu,i} + X_{4,i} C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \\
& - \frac{0.94}{100} Q_{a,i} C_{ta,i} X_{s,i} - \frac{0.94}{100} Q_{a,i} C_{ta,i} X_{p,i} \\
& \leq Q_{I,i} (C_{tp,i} - TDS_{std}) - Q_{a,i} C_{ta,i} - Q_{se,i} C_{tse,i} - Q_{Y,i} C_{ty,i}
\end{aligned} \tag{45}$$

$$X_{1,i} + X_{2,i} - X_{7,i} = 0 \quad i = 1, 2, 3, 4 \tag{46}$$

The computer output of a problem typical of those solved in this report is presented in the appendix. It is apparent that the arithmetic involved in setting up a separable programming problem for the revised simplex method is very cumbersome. For this reason, a

FORTRAN program that was developed by Williams (1972) has been used. The program does all the necessary calculations to linearize the non-linear separable functions of the problem, and produces punched input data in MPS/360 format. To this punched output are added the following cards:

- 1) Cards defining the objective functions and the constraints.
- 2) Cards for the linear portions of the separable programming problem.
- 3) Cards defining the right hand sides of the constraints.
- 4) The necessary control cards.

In solving the problems all the nonlinear separable functions were linearized over 10 partitions.

3.6. OPTIMAL POLICIES FOR THE DISTRICT MODEL

In this section the water quality control policies in each district in the Utah Lake drainage area are considered. At present the high concentration of TDS poses the most serious water quality problem in the area. However with increasing urbanization in the area, the question of municipal and industrial wastes is sure to cause some concern, if not now, in the near future. Thus, in regions like the Utah Lake drainage area, where both urban and agricultural activities prevail side by side, comprehensive water quality management programs must be developed that focus on coordination between urban and agricultural pollution control on a large scale.

In this section, coordinated salinity-BOD control policies for each district are determined under present conditions. In the solutions obtained, it has been assumed that each district is to satisfy the water quality standards individually. Such a policy will be referred to as a 'district policy' in further considerations.

It may be recalled that the optimization problem in separable programming form is to minimize the objective function, $Y_{D,i}$, represented by eq. (41) subject to the constraints as given by eqs. (42) through (46) respectively.

The values of the parameters in these equations, $C_{s,i}$ $C_{p,i}$, the various flow rates, and the influent salt and BOD concentrations are given in Tables 4 and 5.

At the present time, a BOD standard of 25 mg/l in the district's effluents is in force (Utah State Health Division, 1974), therefore this standard is used in each district. Since there are no specific salt concentration standards as of now, the value of the salt reduction standard (TDS), was artificially varied over a wide range to investigate the resulting changes in optimal policies.

Separable programming is used as the method of optimization in each district. The reader is referred to Williams (1972) for details about the computational procedures for the method. A typical computer output is attached in Appendix A.

In Figs. 7 through 10, the total annual costs as well as the sector costs of salinity-BOD control are presented for districts 1 through 4, respectively. It might be observed that all the districts

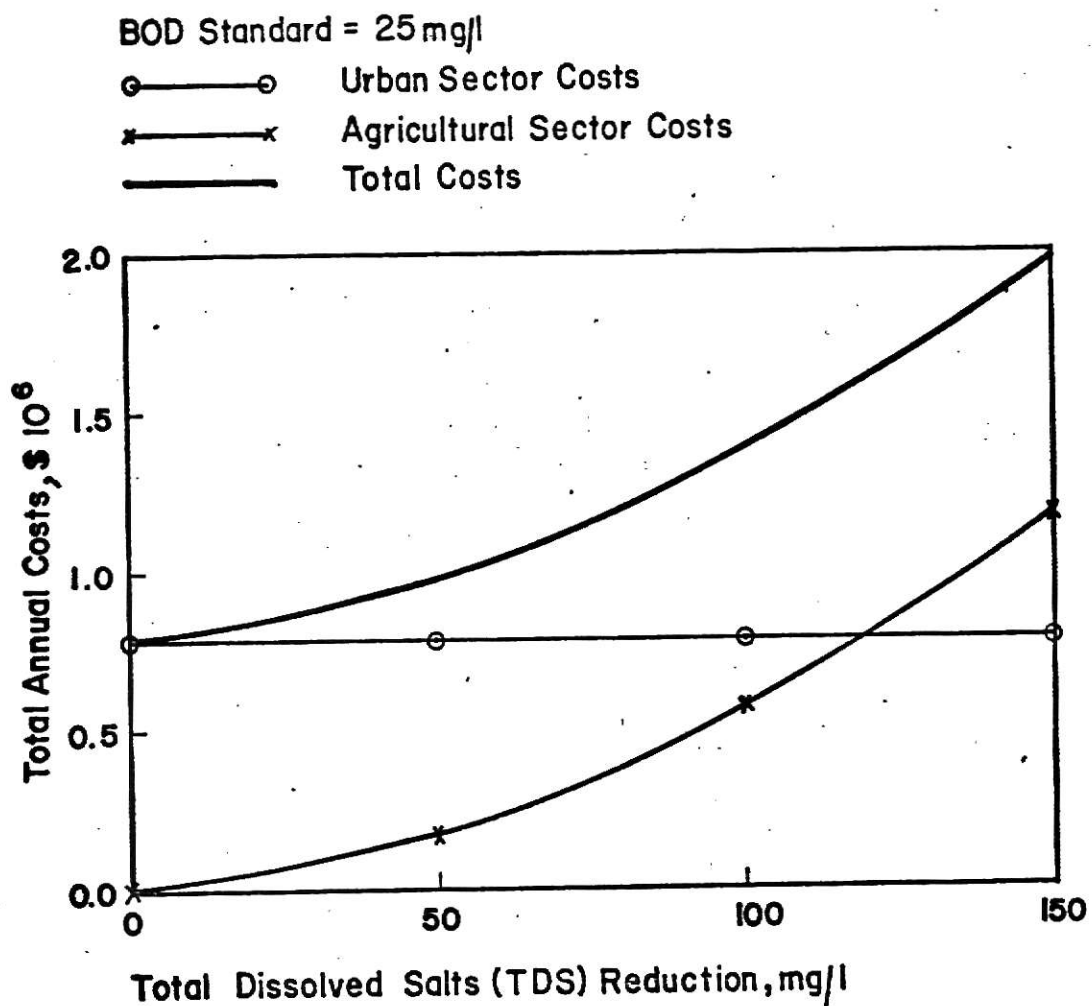


Fig. 7. Total annual costs vs. TDS reduction in the Lehi-American Fork District.

BOD Standard = 25 mg/l

39

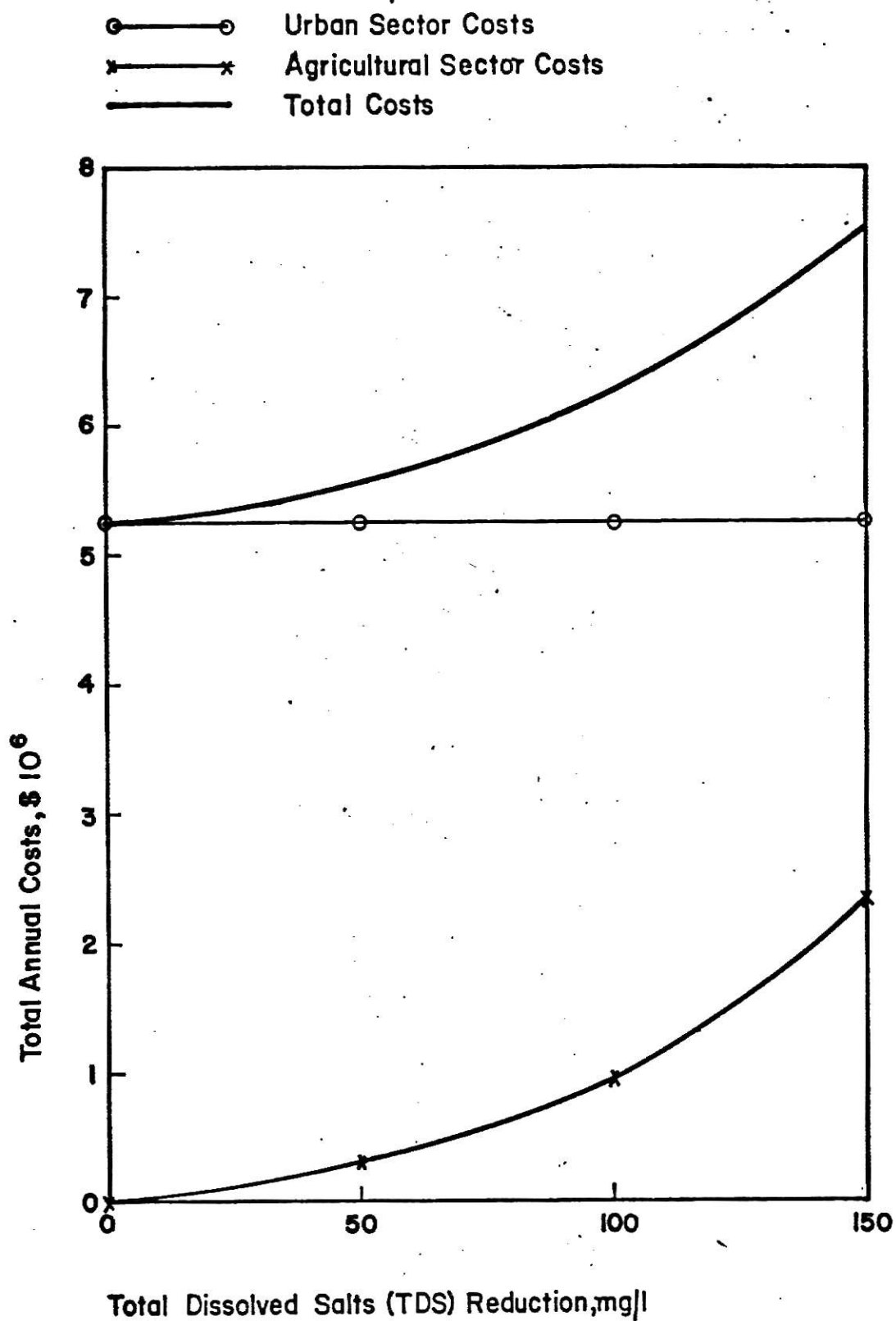


Fig. 8. Total annual costs vs. TDS reduction in the Provo District.

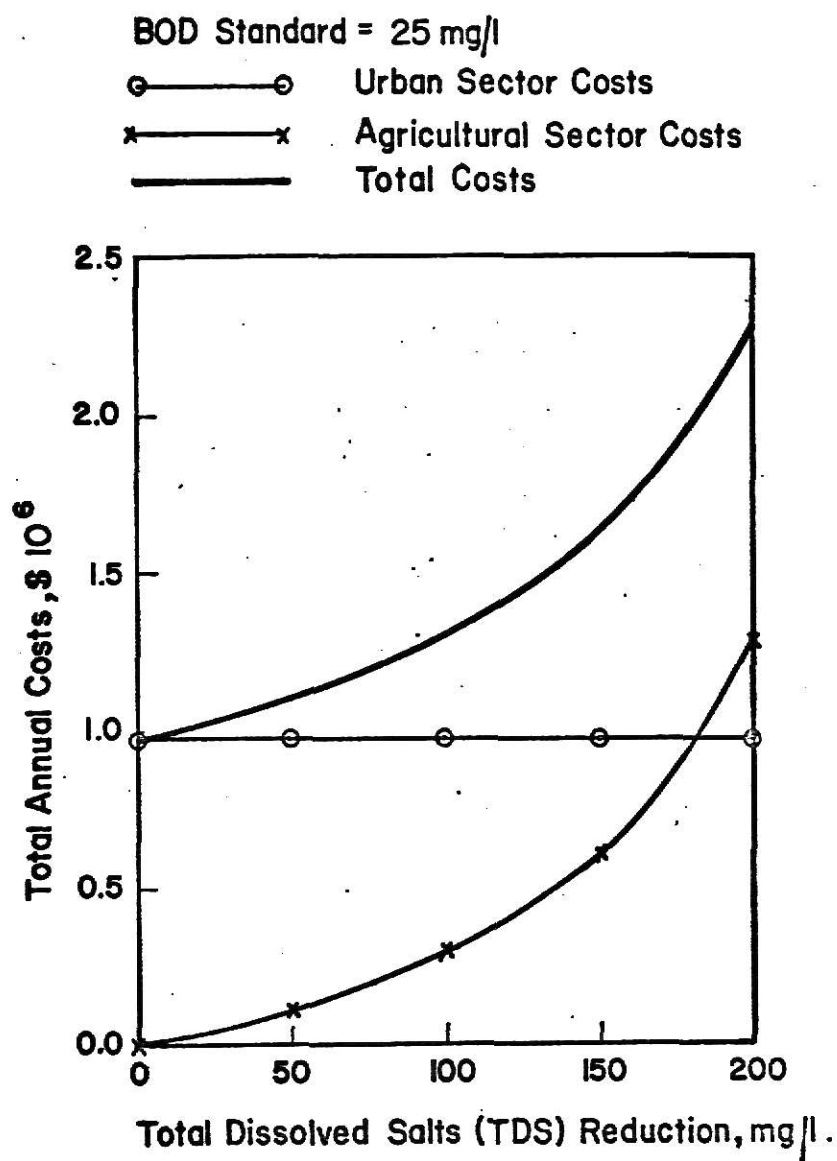


Fig. 9. Total annual costs vs. TDS reduction in the Spanish Fork District.

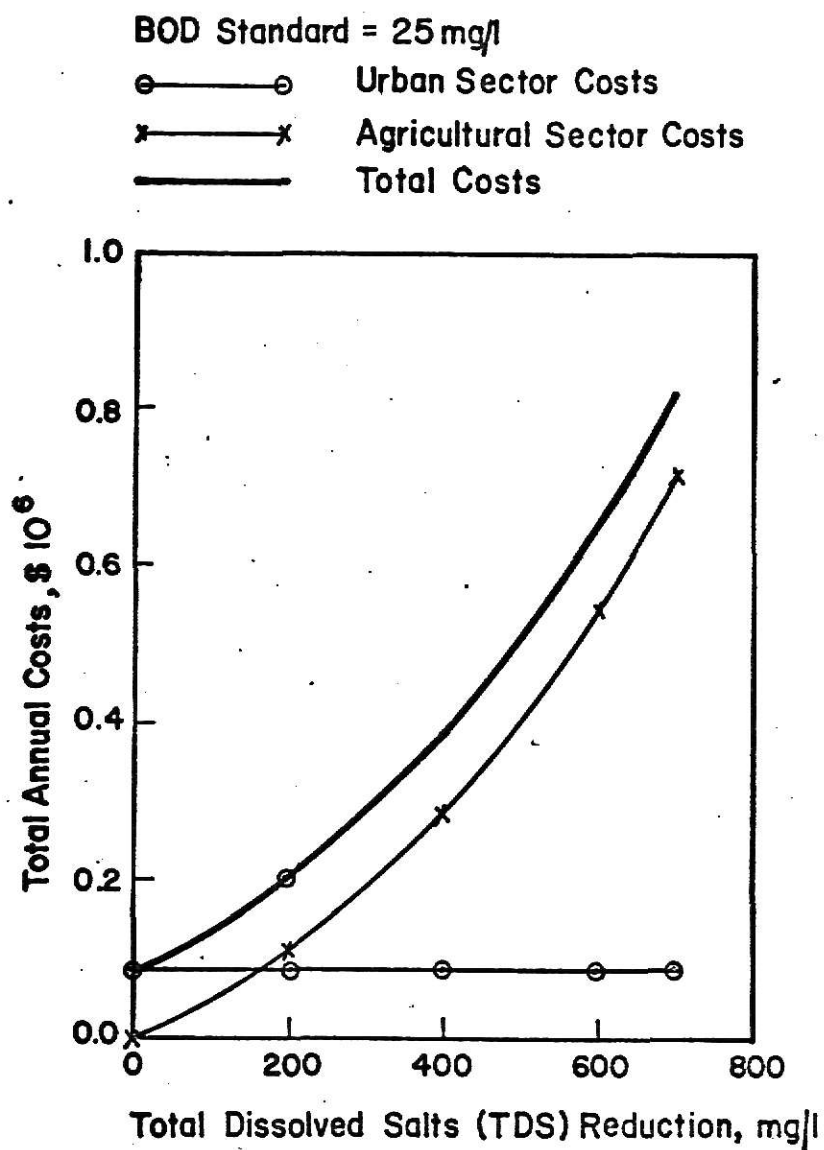


Fig. 10. Total annual costs vs. TDS reduction in the Elberta-Goshen District.

exhibit an increasing marginal cost, implying that low levels of pollution control are cheaper to achieve than higher levels. As an illustration, the cost of increasing the salt reduction level in the district's effluents from 0 mg/l to 50 mg/l is \$0.19 million per year, in the Lehi-American Fork district. On the other hand, the cost for a 50 mg/l increase in salt reduction from 100 mg/l to 150 mg/l is \$0.60 million. There is also a significant difference in water quality control costs between the Lehi-American Fork district and the Provo district. It costs \$1.981 million in the Lehi-American Fork district to maintain a BOD level of 25 mg/l and a TDS reduction standard of 150 mg/l. For the same quality levels, the annual cost in the Provo district is \$7.526 millions. This is attributed to the fact that both urban and agricultural return flows in the Provo district are considerably larger than those of the Lehi-American Fork district. The Spanish Fork district is characterized by a high level of agricultural activity with a low level of urbanization. A BOD standard of 25 mg/l and a salt reduction of 200 mg/l is achieved at a cost of \$2.265 millions, of which \$0.971 million is for the urban sector, and \$1.294 millions is for the agricultural sector. The agricultural return flows in the Elberta-Goshen district while being modest in comparison with those in the other districts, are almost 100 times larger than the urban return flows in the district. The salt content of these return flows is relatively high, and hence a more stringent salt reduction standard should be imposed on the district's effluents. The total annual cost of achieving a BOD standard of 25 mg/l and a salt reduction of 700 mg/l is \$0.823 million.

Tables 6 through 9 give the optimal flows, X_1, X_2, X_3 and X_4 in the urban sector, and the agricultural control measures, X_s and X_p , and also the total annual cost of maintaining the water quality level. The figures correspond to a BOD standard of 25 mg/l. The salt reduction standard is varied over a suitable range for each district.

At this stage, referring back to Fig. 4, it may be noted that the urban return flows are treated in the primary and secondary units, regardless of the BOD standard imposed on the system. If treatment in the primary and secondary units is not sufficient to satisfy the BOD standard in the district, as much of the urban return flows as necessary may undergo tertiary treatment. However, as is evident from the urban sector flow mixture in Tables 6 through 9, primary and secondary treatment of the urban return flows is quite sufficient to reduce the BOD content to such a level, that the BOD standard in the district's effluents is met; hence no tertiary treatment is required in any district.

A second feature of the district policies, as is evident from Tables 6 through 9, is that in all the districts the salt reduction is carried out entirely in the agricultural sector. The implication is that, physical removal of salt through the desalting plant is quite expensive, especially as any flow passing into the desalting plant must undergo tertiary treatment first. Besides, as was mentioned earlier, costs of salinity control in the agricultural sector are lower at low levels of control than corresponding costs in the urban sector, thus making investment in the agricultural sector more effective.

Table 6. Optimal policies for the Lehi-American District, corresponding to a BOD standard of 25 mg/l.

TDS Reduction Standard (mg/l)	Urban Sector Flow Mixture			Agricultural Sector's Control Measures		Total Annual Costs \$10 ⁶
	X ₁ mgd	X ₂ mgd	X ₃ mgd	X ₄ mgd	X _s %	X _p %
0	0	6.42	0	0	0	0.791
50	0	6.42	0	0	21.48	0.980
100	0	6.42	0	0	30.00	1.376
150	0	6.42	0	0	33.75	1.981

Table 7. Optimal Policies for the Provo District, corresponding to a BOD standard of 25 mg/l.

TDS Reduction Standard (mg/l)	Urban Sector Flow Mixture				Agricultural Sector's Control Measures		Total Annual Costs \$10 ⁶
	X ₁ mgd	X ₂ mgd	X ₃ mgd	X ₄ mgd	X _s %	X _p %	
0	0	70.51	0	0	0	0	5.236
50	0	70.51	0	0	25.86	0	5.569
100	0	70.51	0	0	30.00	21.73	6.226
150	0	70.51	0	0	33.75	43.85	7.526

Table 8. Optimal Policies for the Spanish Fork District, corresponding to a BOD standard of 25 mg/l.

TDS Reduction Standard (mg/l)	Urban Sector Flow Mixture				Agricultural Sector's Control Measures		Total Annual Costs \$10 ⁶
	X ₁ mgd	X ₂ mgd	X ₃ mgd	X ₄ mgd	X _s %	X _p %	
0	0	6.71	0	0	0	0	0.971
50	0	6.71	0	0	10.41	0	1.107
100	0	6.71	0	0	20.79	0	1.305
150	0	6.71	0	0	30.00	1.18	1.667
200	0	6.71	0	0	33.75	7.82	2.265

Table 9. Optimal Policies for the Elberta-Goshen District, corresponding to a BOD standard of 25 mg/l.

TDS Reduction Standard (mg/l)	Urban Sector Flow Mixture				Agricultural Sector's Control Measures		Total Annual Costs \$10 ⁶
	X ₁ mgd	X ₂ mgd	X ₃ mgs	X ₄ mgd	X _s %	X _p %	
0	0	0.20	0	0	0	0	0.092
150	0	0.20	0	0	14.70	0	0.176
200	0	0.20	0	0	17.08	0	0.209
400	0	0.20	0	0	26.50	7.84	0.385
600	0	0.20	0	0	30.00	21.10	0.648
700	0	0.20	0	0	31.61	28.00	0.823

In the discussion in this section, it has been assumed that the cost of zero mg/l salt reduction is zero. Referring back to the objective function, represented by eq. (25), it is noted that the term

$$C_{p,i} \times 37.60 [4.11087 - n (56.50 - X_{p,i})]/100$$

does not reduce to zero when $X_{p,i}$ is zero. For the Lehi-American Fork district, the above expression reduces to \$0.073 million when $X_{p,i} = 0$. Corresponding values for the other districts are \$0.086 million, \$0.226 million and \$0.049 million, for the Provo, Spanish Fork and Elberta-Goshen districts, respectively. These values have been added to the urban sector costs in presenting the results.

3.7. COMPARISON OF THE SEPARABLE PROGRAMMING SOLUTION WITH THE GENERALIZED REDUCED GRADIENT METHOD SOLUTION

The same problem for the Utah Lake drainage area has been solved by Shojalashkari (1974). In this section, some of the results obtained using separable programming are compared with the results obtained using the G.R.G. method. Table 10 presents such a comparison, for a BOD standard of 25 mg/l and a TDS reduction standard of 150 mg/l. Referring to Table 10, it is observed that the two solutions are quite close to each other. The slight differences may be attributed to the fact that separable programming does involve some approximation in linearizing the nonlinear separable functions.

Table 10. Comparison of the Separable Programming solution with the G.R.G. solution.

For all districts BOD standard = 25 mg/l
 TDS reduction standard = 150 mg/l

District	Method of Solution	Urban Sector Flow Mixture (mgd)				Agricultural Sector Control Measures %			Total Annual Costs \$10 ⁶
		X ₁	X ₂	X ₃	X ₄	X _s	X _p		
1	Separable Prog.	0	6.42	0	0	33.75	30.04	1.981	
	G.R.G.	0	6.42	0	0	33.10	30.40	1.950	
2	Separable Prog.	0	70.51	0	0	30.00	21.73	7.526	
	G.R.G.	0	70.51	0	0	35.70	42.10	7.440	
3	Separable Prog.	0	6.71	0	0	30.00	1.18	1.667	
	G.R.G.	0	6.71	0	0	29.20	1.50	1.700	
4	Separable Prog.	0	0.20	0	0	14.70	0	0.176	
	G.R.G.	0	0.20	0	0	13.00	0	0.160	

3.8. REVIEW OF A REPORT BY FUHRIMAN, MERRITT, ET AL. (1975)

In this section a brief review of a report by Fuhriman, Merritt, et al., (1975) is given. In their work, the inflows, outflows, and in-lake water quality and quantity of the Utah Lake were studied over a 36 month period in order to determine the effect of a proposed diking project on the quality and quantity of the water in the Lake. Also, a methodology was developed to determine the effects of diking, or for that matter, any other water management practice on the quality of water in a lake system.

The report also presents a computer simulation model that analyzes the effects of a water management program on the water quality of the Lake, particularly as related to the 'conservative salts' present. The simulation model was also used to evaluate water evaporation from the Lake by use of a salt balance technique. The results obtained in the study confirmed that diking of the Utah Lake would have positive beneficial effects on the water quality in the Lake. Diking will also result in considerable saving of water, and reclamation of valuable land.

As was mentioned earlier, the work involved measuring inflows and outflows into the Lake, as well as water quality of the flows. A comparison of some of the results obtained in this study with those of the report was attempted. In fact, the report does mention that quality and quantity measurements on the effluents from the Lehi-American Fork and Provo sewage treatment plants were taken. But the computer listing that followed did not contain any information on these effluents.

CHAPTER 4

THE REGIONAL SALINITY-BOD CONTROL MODEL

In the previous chapter, optimal policies were developed under the assumption that each district operates independently. Each district is required to satisfy the BOD and TDS reduction standards in its effluents. It is worth investigating whether it would be more economical for the districts to coordinate their salinity-BOD control activities, and operate on a regional basis. Under this mode of operation the BOD and TDS reduction standards would be met in the aggregated effluents into the Lake, though they may be violated in some individual district's effluents.

Advantages of a regional approach also stem from the economies of operating on a larger scale. Problems faced by local governments, which lack the necessary economic and manpower resources, can be alleviated (Clayton and Huie, 1973).

In this chapter, an attempt is made to explore the potentials for regionalization and coordination of the water control strategies among the four districts of the Utah Valley.

4.1. FORMULATION OF THE REGIONAL SALINITY-BOD CONTROL MODEL

The objective in the formulation of the Utah Valley salinity-BOD model, is to minimize the aggregated costs of urban wastewater treatment as well as the costs of salinity control in the districts, such that the BOD and TDS reduction standards in the aggregated inflows into the Utah Lake are met.

It must be stressed, that in the regional model, there are still four waste water treatment facilities in operation, one in each district. One might wonder, why not have a single waste water treatment facility for the entire Utah Valley drainage area. Consolidation of the smaller waste water treatment plants into a regional facility might be supported by arguments based on the economies of operating on a larger scale, that were mentioned earlier. However, the cost reduction achieved by operating and maintaining such a facility, should be weighed against the cost of transporting the waste water to this central facility. Optimal location of the centralized facility itself would be a problem in operations research. As information on the cost of waste water transportation in the Utah Valley drainage area was not available the existing structure of urban waste water treatment was left unchanged. In considerations that follow, each district still operates its own waste water treatment facility.

Under the above assumption, the objective function of the Utah Valley model can be written as follows:

Minimize

$$Y_T = \sum_{i=1}^4 Y_{D,i}$$

where

Y_T represents the total annual cost of salinity-BOD control in the Utah Valley area.

$Y_{D,i}$ is the total annual cost of salinity-BOD control in the i district, as given by eq. (25).

As before the constraints on the system may be classified into flow constraints, and BOD and TDS reduction constraints. It might be recalled, that each district still operates its own urban waste water treatment facility. Hence the flow constraints must be satisfied in each district individually. The flow constraints may be stated as follows:

$$g_{1,i}(\bar{X}) = 0 \quad i = 1,2,3,4 \quad (47)$$

$$g_{2,i}(\bar{X}) = 0 \quad i = 1,2,3,4 \quad (48)$$

where $g_{1,i}(\bar{X})$ and $g_{2,i}(\bar{X})$ are given by eqs. (26) and (27), respectively.

The BOD constraints on the urban and agricultural return flows are imposed on each district effluent, individually. It might seem to the reader, that this is contradictory to the purpose of coordinating water quality control strategies among the districts. But legal requirements about the minimum amount of BOD reduction that must be carried out by the polluters, force imposing the BOD constraints on each district individually. Therefore the BOD constraints can be written as follows:

$$g_{3,i}(\bar{X}) \leq 0 \quad , i = 1,2,3,4 \quad (49)$$

where $g_{3,i}(\bar{X})$ is given by eq. (35).

The TDS reduction standard can be imposed on the aggregated return flows into the Lake. The constraint can be written as:

$$\begin{aligned}
g_6(\bar{X}): \sum_{i=1}^4 \left\{ (X_{2,i} + X_{4,i}) C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \right. \\
+ Q_{a,i} C_{ta,i} - \frac{0.94}{100} C_{ta,i} (X_{s,i} + X_{p,i}) \\
+ Q_{se,i} C_{tse,i} + Q_{y,i} C_{ty,i} \left. \right\} \\
- \left\{ \sum_{i=1}^4 (Q_{u,i} + Q_{a,i} + Q_{se,i} + Q_{y,i} - Q_{w,i}) \right\} \\
\times (C_{tpe} - TDS_{rtd}) \leq 0 \quad (50)
\end{aligned}$$

where C_{tpe} is the present salt content of the aggregated return flows, computed as follows:

$$\frac{\sum_{i=1}^4 (Q_{u,i} C_{tu,i} + Q_{a,i} C_{ta,i} + Q_{y,i} C_{ty,i} + Q_{se,i} C_{tse,i})}{\sum_{i=1}^4 (Q_{u,i} + Q_{a,i} + Q_{y,i} + Q_{se,i} - Q_{w,i})} \quad (51)$$

TDS_{std} is the salt reduction standard.

The optimization problem may be summarized as:

Minimize

$$Y_T = \sum_{i=1}^4 Y_{D,i} \quad (52)$$

subject to

$$g_{1,i}(\bar{X}) = 0 \quad i = 1, 2, 3, 4$$

$$g_{2,i}(\bar{X}) = 0 \quad i = 1, 2, 3, 4$$

$$g_{3,i}(\bar{X}) \leq 0 \quad i = 1, 2, 3, 4$$

$$g_6(\bar{X}) \leq 0$$

where

$Y_{D,i}$ is defined by eq. (25), $g_{1,i}(\bar{X})$, $g_{2,i}(\bar{X})$, $g_{3,i}(\bar{X})$ and $g_6(\bar{X})$ are defined by eqs. (26), (27), (35) and (50), respectively.

4.2. ADAPTING THE REGIONAL MODEL TO SEPARABLE PROGRAMMING

Referring back to the mathematical representation of the Regional Model, it is observed that only the objective function contains nonlinear terms that are not separable.

Once again the following simple substitutions yield an objective function that is made up of separable functions:

$$X_{1,i} + X_{2,i} = X_{7,i} \quad , i = 1, 2, 3, 4$$

The optimization problem in separable programming form may be written as follows:

Minimize

$$\begin{aligned} Y_T = \sum_{i=1}^4 \{ & 3.31 \times 10^4 X_{7,i}^{0.71} + 2.12 \times 10^4 X_{7,i}^{0.83} + 6.07 \times 10^4 X_{7,i}^{0.80} \\ & + 4.29 \times 10^4 X_{7,i}^{0.893} + 5.23 \times 10^3 X_{1,i}^{0.89} + 4.16 \times 10^4 X_{1,i}^{0.65} \\ & + 4.16 \times 10^3 X_{1,i}^{0.90} + 1.42 \times 10^4 X_{1,i}^{0.962} + 4.75 \times 10^4 X_{1,i}^{0.72} \\ & + \begin{cases} 5.50 \times 10^4 X_{1,i}^{0.70} & , X_{1,i} \leq 3 \text{ mgd} \\ 5.70 \times 10^3 X_{1,i}^{0.96} & , X_{1,i} > 3 \text{ mgd} \end{cases} \\ & + 5.34 \times 10^4 X_{3,i}^{0.67} + 22.76 \times 10^4 X_{3,i}^{0.79} \end{aligned}$$

$$\begin{aligned}
& + C_{s,i} \times 34.42 \left[3.68135 - \ln (39.70 - X_{s,i}) \right] / 100 \\
& + C_{p,i} \times 37.60 \left[4.11087 - \ln (56.50 - X_{p,i}) \right] / 100 \} \quad (53)
\end{aligned}$$

subject to

$$X_{1,i} + X_{2,i} = Q_{u,i} \quad , i = 1, 2, 3, 4 \quad (54)$$

$$X_{2,i} + X_{3,i} + X_{4,i} = Q_{u,i} \quad , i = 1, 2, 3, 4 \quad (55)$$

$$\begin{aligned}
& X_{2,i} (1-e_{ps}) C_{bu,i} + X_{3,i} (1-e_t)(1-e_{ps})C_{bu,i} \\
& + X_{4,i} (1-e_t)(1-e_{ps})C_{bu,i} \\
& \leq (Q_{u,i} + Q_{a,i}) \times BOD_{std} - Q_{a,i} C_{ba,i} \quad , i = 1, 2, 3, 4 \quad (56)
\end{aligned}$$

$$\begin{aligned}
& \sum_{i=1}^4 \left\{ (X_{2,i} + X_{4,i}) C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \right. \\
& \left. - \frac{0.94}{100} Q_{a,i} C_{ta,i} X_{s,i} - \frac{0.94}{100} Q_{a,i} C_{ta,i} X_{p,i} \right\} \\
& \leq - \sum_{i=1}^4 \left\{ Q_{a,i} C_{ta,i} + Q_{se,i} C_{tse,i} + Q_{y,i} C_{ty,i} \right\} \\
& - \left\{ \sum_{i=1}^4 (Q_{u,i} + Q_{a,i} + Q_{se,i} + Q_{y,i} - Q_{w,i}) \right\} \\
& \quad \times (C_{tpe} - TDS_{std}) \quad (57)
\end{aligned}$$

$$X_{1,i} + X_{2,i} - X_{7,i} = 0 \quad , i = 1, 2, 3, 4 \quad (58)$$

The computer solution of the problem has already been dealt with in a previous chapter and will not be dealt with here.

4.3. OPTIMAL COSTS AND POLICIES FOR THE REGIONAL MODEL

The optimization problem for Utah Valley in separable programming form, as represented by the objective function, eq. (53), and the constraint, eqs. (54) through (58), has 28 variables, 4 of which are dummy variables introduced to adapt the problem to separable programming. The problem has 17 constraints, all of which are linear. The separable programming option in MPS/360 is once again used to solve the problem. A BOD standard of 25 mg/l and a TDS reduction standard of 150 mg/l are assumed. The resulting optimal policies are presented in Table 10.

The urban sector flow mixtures in Table 10, indicate that the urban return flows, after undergoing primary and secondary treatment, are entirely diverted to the urban effluent channels. The implication is that no tertiary treatment is required to meet the BOD standard. In fact, substitution of the results in the BOD constraints would reveal that the BOD constraints are far from active. They are very easily satisfied by primary and secondary treatment of the urban return flows. Once again, as in the case of independent or individual operation, salt reduction is carried out entirely in the agricultural sector. However, the agricultural control measures, $X_{s,i}$ and $X_{p,i}$, are distributed among the districts in a different fashion from when the districts were operating independently. For the regional model, structural improvements and practice improvements in the Lehi-American Fork district are 26.25% and zero % (see Table 11), respectively, as against 33.75% and 30.04% (see Table 10) for independent operation of

the districts. For the Provo district, the corresponding values of X_s and X_p are 30.00% and 16.80% (see Table 11) for the regional model, as against 33.75% and 43.85% (see Table 10) for independent operation; for the Spanish Fork district, the values for the regional model are 30.00% and 5.60% (see Table 11) as against 30.00% and 1.18% (see Table 10) for independent operation; for the Elberta-Goshen district, the values are 26.25% and 11.20% (see Table 11) for the regional model, as against 14.70% and zero % (see Table 10) for independent operation. It is noted that for regional operation, control measures in the agricultural sector are reduced for the Lehi-American Fork and Provo districts, while they are augmented in the other two districts.

The economic advantages of a regional strategy in comparison with individual operation, will now be dealt with. In Table 11, the total annual cost for the Utah Valley, as well as the cost breakdown for the districts are given. For a BOD standard of 25 mg/l and a TDS reduction standard of 150 mg/l, the total annual cost for regional operation is $\$9.522 \times 10^6$, the breakdown among the districts being as follows:

$\$1.054 \times 10^6$ for the Lehi-American Fork district,
 $\$6.076 \times 10^6$ for the Provo district,
 $\$1.963 \times 10^6$ for the Spanish Fork district,
 $\$0.429 \times 10^6$ for the Elberta-Goshen district.

The total annual cost under the assumption of individual operation are $\$11.346 \times 10^6$, obtained by summing up the following:

Table 11. Optimal policies for the Utah Valley Drainage Area, corresponding to a BOD standard of 25 mg/l and a salt reduction standard of 150 mg/l

District	Urban Sector Flow Mixture				Agricultural Sector Control Measures		Total Annual Cost \$10 ⁶
	X ₁ mgd	X ₂ mgd	X ₃ mgd	X ₄ mgd	X _s %	X _p %	
Lehi-American Fork District	0 (0)*	6.42 (6.42)	0 (0)	0 (0)	26.25 (26.3)	0 (14.6)	
Provo District	0 (0)	70.51 (70.51)	0 (0)	0 (0)	30.00 (31.9)	16.8 (20.6)	9.522
Spanish Fork District	0 (0)	6.71 (6.71)	0 (0)	0 (0)	30.90 (33.0)	5.60 (14.2)	(10.465)
Elberta-Goshen District	0 (0)	0.20 (0.20)	0 (0)	0 (0)	26.25 (26.3)	11.20 (0)	

* Figures in parentheses indicate optimal policies obtained using the Generalized Reduced Gradient Method by Shojalashkari.

Table 12. Cost breakdown for the Utah Valley Model. Figures in \$10⁶

Cost Breakdown	Districts			
	Lehi-American Fork	Provo	Spanish Fork	Elberta-Goshen
1) Urban Sector	0.791	5.236	0.971	0.092
2) Agricultural Sector				
Structural Improvements	0.263	0.444	0.684	0.196
Practice Improvements	0.000	0.396	0.308	0.141
Total Agricultural Sector	0.263	0.840	0.992	0.337
3) Total Cost for District	1.054	6.076	1.963	0.429
4) Total Cost for Utah Valley		9.522		

$\$1.981 \times 10^6$ for the Lehi-American Fork district,

$\$7.526 \times 10^6$ for the Provo district,

$\$1.667 \times 10^6$ for the Spanish Fork district,

$\$0.176 \times 10^6$ for the Elberta-Goshen district.

Thus regional operation results in a cost reduction of $\$1.824 \times 10^6$ per year as compared with independent operation.

In summary, it may be pointed out that the significant savings resulting from adopting a regional approach, stems from the fact that each district does not have to meet the salt reduction standards individually. There is reallocation of salt reduction measures among the agricultural sectors of the four districts. In case of the Lehi-American Fork district and the Provo district, agricultural control measures and consequently, costs, are less under a regional mode of operation. On the other hand, agricultural salinity control measures are augmented in case of the Spanish Fork district and the Elberta-Goshen district. Costs in these two districts are correspondingly higher than under the assumption of independent operation. The urban waste water treatment costs remain the same for both regional and independent operation. This is because even under the assumption of regional operation, the BOD constraints are imposed individually on the return flows of each district. Besides, the construction of the model requires that, in each district the urban return flows undergo primary and secondary treatment in any case.

4.4. COMPARISON OF THE SEPARABLE PROGRAMMING SOLUTION WITH THE GENERALIZED REDUCED GRADIENT METHOD SOLUTION

The figures indicated within parentheses in Table 11 are the results obtained by Shojalashkari for the same problem using the Generalized Reduced Gradient method. It is noted that the two methods yield identical urban sector policies. There are some differences in the optimal policies for salinity control in the agricultural sectors. The most significant differences are the following:

The optimal policy obtained using separable programming suggests a control measure of $X_{p,i}$ = zero, in practice improvements in the Lehi-American Fork district, and 11.2% in practice improvements in the Elberta-Goshen district. The corresponding values in the G.R.G. solution obtained by Shojalashkari are 14.6% in the Lehi-American Fork district, and zero % in the Elberta-Goshen district. Practice improvement control measures are 16.8% and 5.60% in the Provo and Spanish Fork districts, respectively, for the separable programming solution. Corresponding values for the G.R.G. solution are 20.6% and 14.2%, respectively. There is also some difference in the total annual cost figures for the two methods of solution; \$9.522 millions for the separable programming solution, as against \$10.465 millions for the G.R.G. solution.

These differences are to be expected in view of the fact that separable programming is a technique that involves approximating nonlinear separable functions by piecewise linear approximations. The closeness of the approximated piecewise linearization to the nonlinear

separable function depends on the shape of the nonlinear curve, as also on the number of partitions used in the above said approximation. Better approximations may be obtained with more flat curves. Increasing the number of partitions also yields a better approximation. However, it must be kept in mind that increasing the number of partitions also increases the number of variables in the problem. Consequently the computer time required in the solution of the problem is also increased.

CHAPTER 5

VERIFICATION OF THE OPTIMAL POLICIES OBTAINED BY WALKER, ET AL. (1973a, 1973b, 1974)

This chapter deals with the discussion of the problem as solved by Walker, et al. (1973b) in their report, and also the results obtained by them. The method of optimization used by Walker, et al. (1973a) has been described as the Jacobian Differential Algorithm.

5.1. PROBLEM AS SOLVED BY WALKER, ET AL. (1973a, 1973b, 1974)

The essentials of the model used by Walker, et al., are the same as that of Shojalashkari (1972). However, Walker, et al., have not included $Q_{y,i}$ (yields from the valley floor), $Q_{se,i}$ (spills and excesses) and $Q_{w,i}$ (wetland consumption), in the final inflow into the Lake. There is a significant salt content associated with these flows. Referring back to Table 7, it is noted that the Lehi-American Fork district has a salt content of 190 mg/l, and the Spanish Fork district has a salt content of 350 mg/l. Besides, the net flow $Q_y + Q_{se} - Q_w$ does contribute a sizable portion of the total inflow into the Lake. For example, in the Lehi-American Fork district, $Q_y + Q_{se} - Q_w$ amounts to 31.5 mgd; total inflow from the district into the Lake is 94.50 mgd. In the Spanish Fork district, it amounts to 49.18 mgd out of a total inflow into the Lake of 158.89 mgd. The situation is somewhat different in the Elberta-Goshen district. $Q_y + Q_{se} - Q_w$ is a negative quantity, -4.05 mgd; total inflow into the Lake is 15.40 mgd. This is because of the rather large magnitude of Q_w .

The optimization problem considered by Walker, et al., can be represented as:

Minimize

$$Y_{D,i}$$

subject to the constraints,

$$g_{1,i}(\bar{X}) = 0$$

$$g_{2,i}(\bar{X}) = 0$$

$$g_{3,i}(\bar{X}) \leq 0$$

$$\begin{aligned} g_{7,i}(\bar{X}) = & (X_{2,i} + X_{4,i}) C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \\ & + Q_{a,i} C_{tu,i} - \frac{0.94}{100} C_{ta,i} (X_{s,i} + X_{p,i}) \\ & - (Q_{u,i} + Q_{a,i}) \times (C_{tp,i} - TDS_{rtd}) \leq 0 \end{aligned} \quad (59)$$

$$i = 1, 2, 3, 4$$

where

$$C_{tp,i} = \frac{(Q_{u,i} \times C_{tu,i} + Q_{a,i} C_{ta,i})}{(Q_{u,i} + Q_{a,i})} \quad (60)$$

and, $Y_{D,i}$, $g_{1,i}(\bar{X})$, $g_{2,i}(\bar{X})$ and $g_{3,i}(\bar{X})$ are given by eqs. (25), (26), (27) and (35).

The regional model for the Utah Valley is formulated as before. The optimization problem is to minimize the aggregated costs, subject to the flow constraints in each district, the BOD constraints in each

district, and the salt reduction constraint on the aggregated return flows into the Lake. In mathematical notation the problem may be defined as:

Minimize

$$Y_T = \sum_{i=1}^4 Y_{D,i}$$

subject to

$$g_{1,i}(\bar{X}) = 0 \quad , i = 1, 2, 3, 4$$

$$g_{2,i}(\bar{X}) = 0 \quad , i = 1, 2, 3, 4$$

$$g_{3,i}(\bar{X}) \leq 0 \quad , i = 1, 2, 3, 4$$

$$\sum_{i=1}^4 g_{7,i}(\bar{X}) \leq 0$$

The last constraint may be expanded as:

$$\begin{aligned} \sum_{i=1}^4 g_{7,i}(\bar{X}) = \sum_{i=1}^4 \left\{ (X_{2,i} + X_{4,i}) C_{tu,i} + (1-e_d) C_{tu,i} X_{3,i} \right. \\ \left. + Q_{a,i} C_{ta,i} - \frac{0.94}{100} C_{ta,i} (X_{s,i} + X_{p,i}) \right\} \\ - \sum_{i=1}^4 (Q_{u,i} + Q_{a,i}) (C_{tpe} - TDS_{rtd}) \leq 0 \end{aligned} \quad (61)$$

where

$$C_{tpe} = \frac{\sum_{i=1}^4 (Q_{u,i} C_{tu,i} + Q_{a,i} C_{ta,i})}{\sum_{i=1}^4 (Q_{u,i} + Q_{a,i})} \quad (62)$$

5.2. A DISCUSSION OF THE OPTIMAL POLICIES OBTAINED BY WALKER, ET AL.

Walker, et al., (1973b) have used the Jacobian Differential Algorithm in the solution of the above optimization problem. The TDS reduction standard was artificially varied over a suitable range, keeping the BOD standard fixed, to study the resulting effect on the optimal policies. In each case, salinity control costs (salinity control cost for a TDS reduction standard of zero mg/l was assumed to be \$0), were plotted against the corresponding TDS reduction standard in mg/l. Plots of TDS reduction in the urban and agricultural sector effluents against TDS reduction in the district's effluents were also constructed. The plots obtained by Walker, et al., have been reproduced in Figs. 11 through 14.

A brief discussion of the results, as presented by Walker, et al., follows. The Provo district is the only one of the four, which, under present conditions, has urban water demands of a magnitude comparable with that of the agricultural water demands. This district contributes about 40% of the return flows in the basin, and 27% of the salts. Referring to Fig. 11, it is observed that, for TDS reductions less than 10 mg/l in the district's effluents, salinity control is primarily an agricultural policy. However, as the necessity for more stringent TDS reduction standards arises, agricultural control measures remain essentially fixed, while urban policies are implemented. Finally, beyond a TDS reduction of 130 mg/l, the balance again tilts in favor of agricultural control measures. The cost curve reflects the alternating policies, which indicate increasing and then decreasing

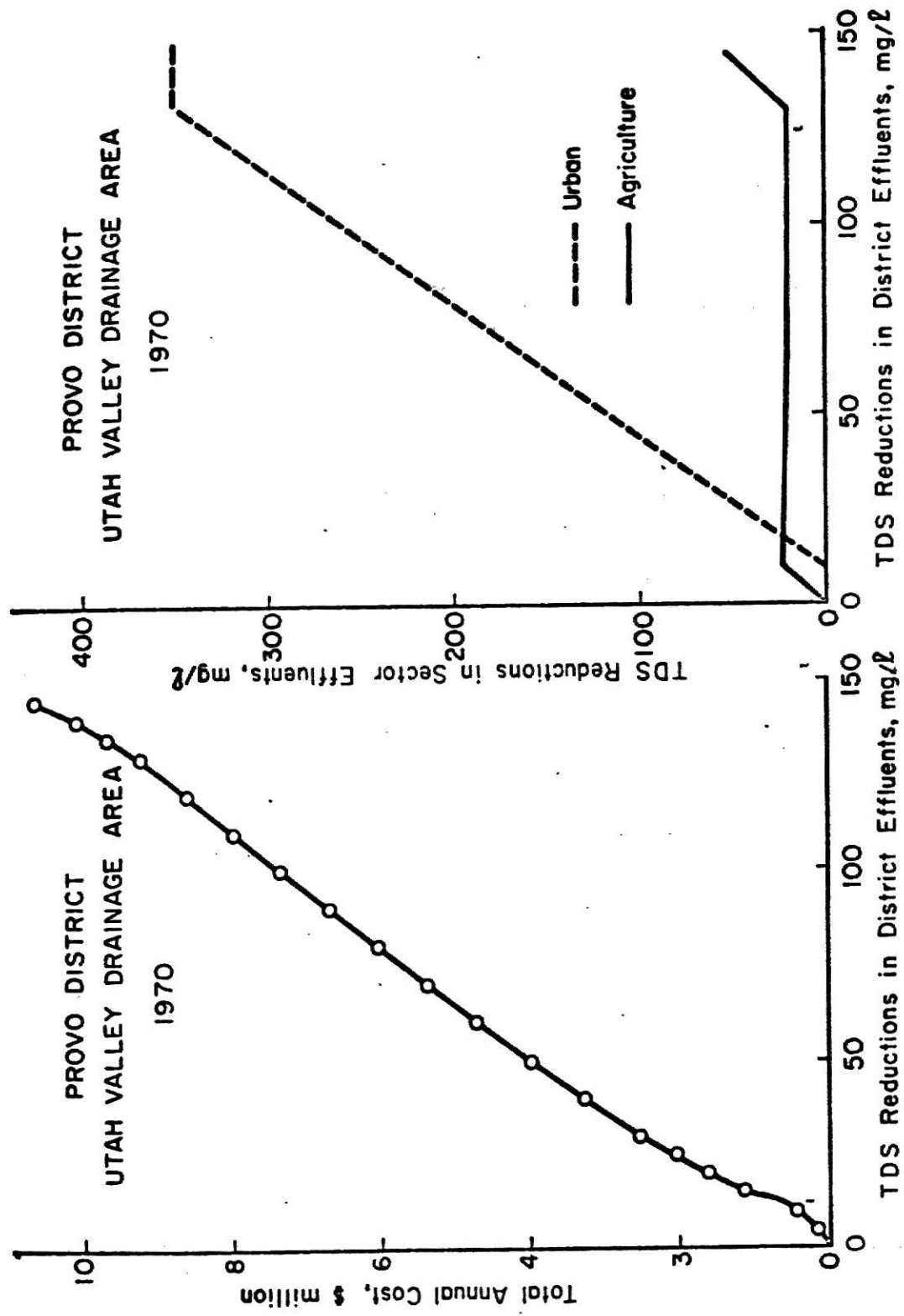


Fig. 11. Optimal policies for salinity control in the Provo District (Walker, et al., 1973b).

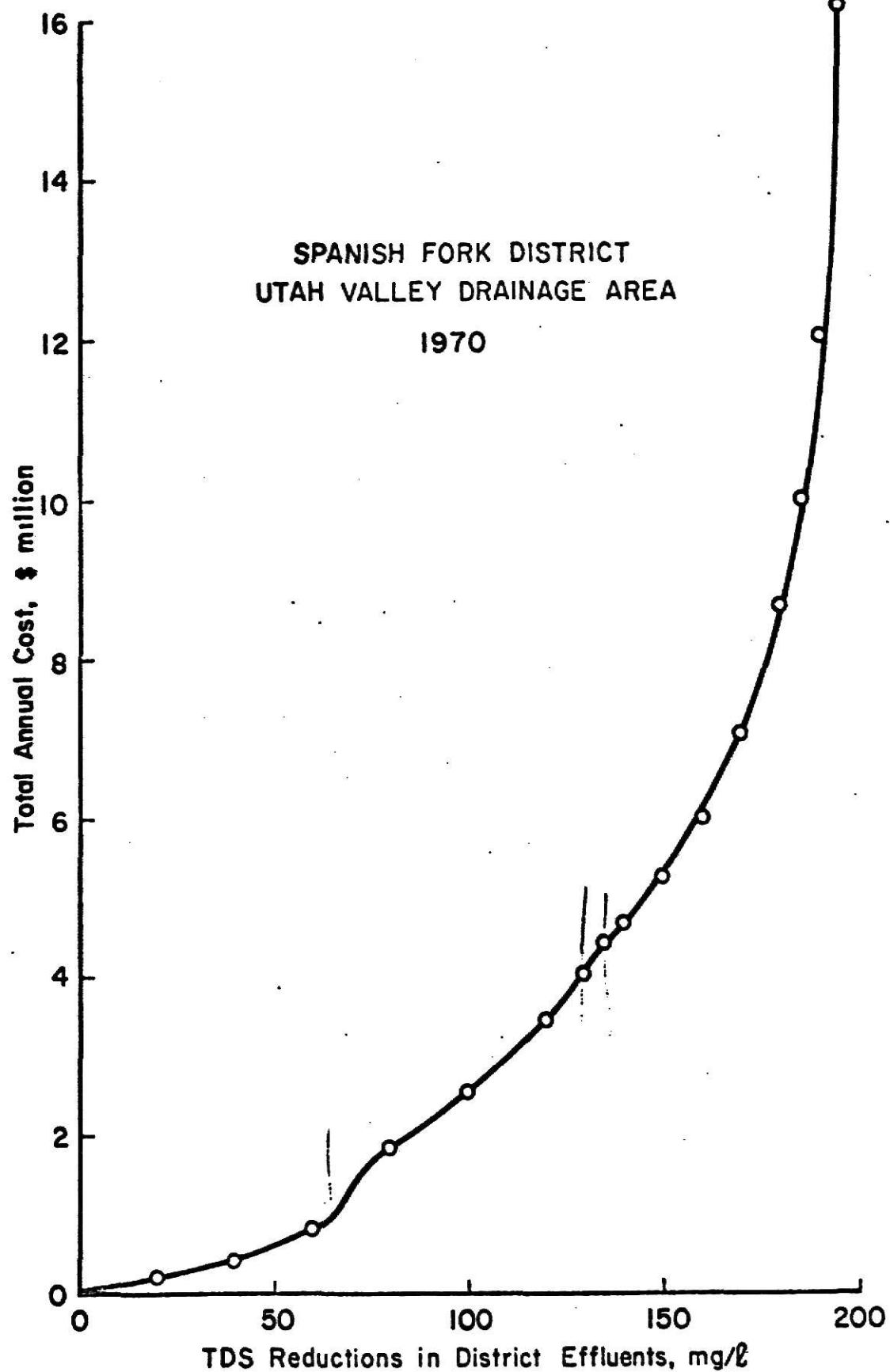


Fig. 12. Total annual costs for salinity control in the Spanish Fork District (Walker, et al., 1973b).

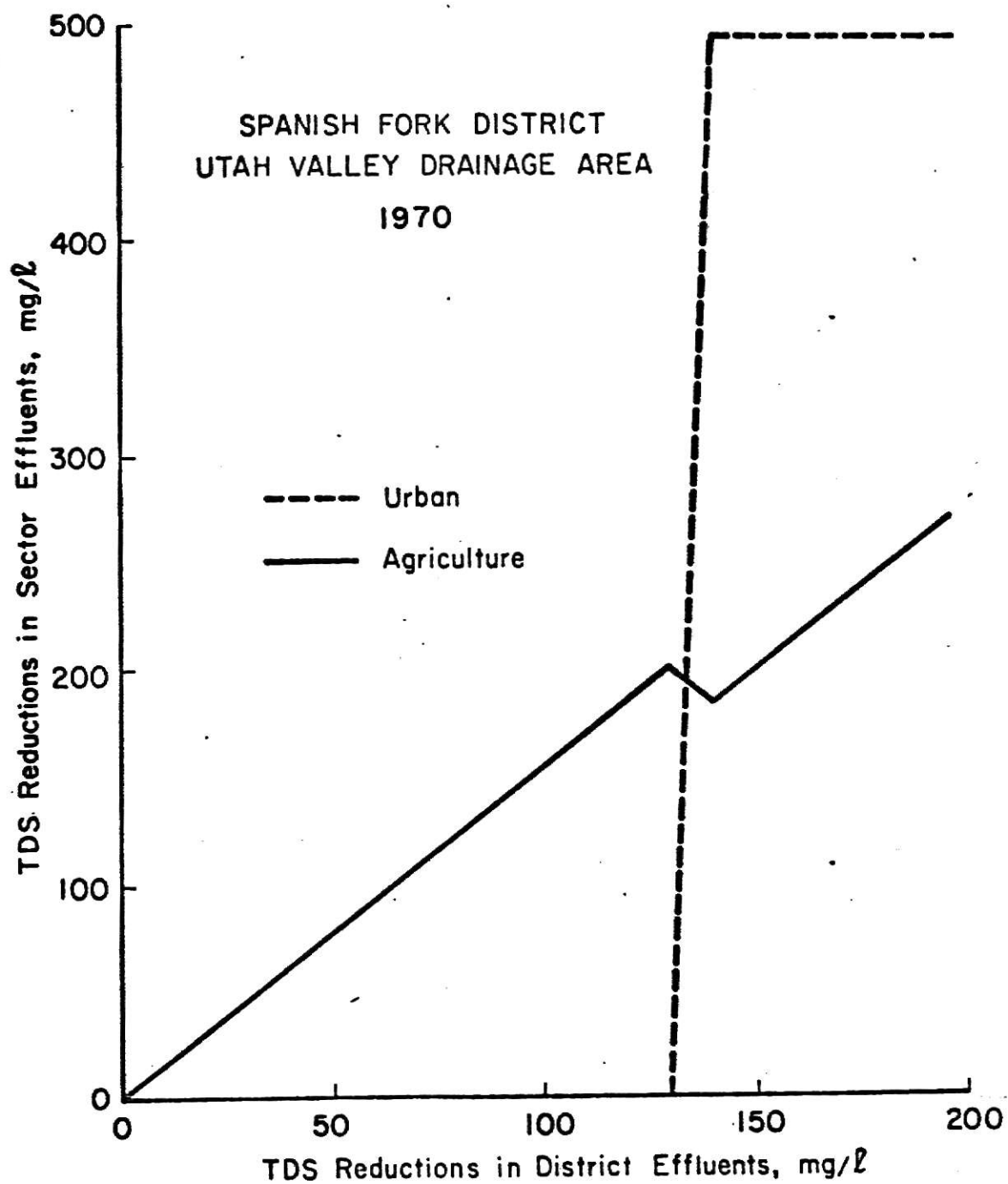


Fig. 13. Optimal policy for reducing the concentrations of TDS in the return flows of the Spanish Fork District (Walker, et al., 1973b).

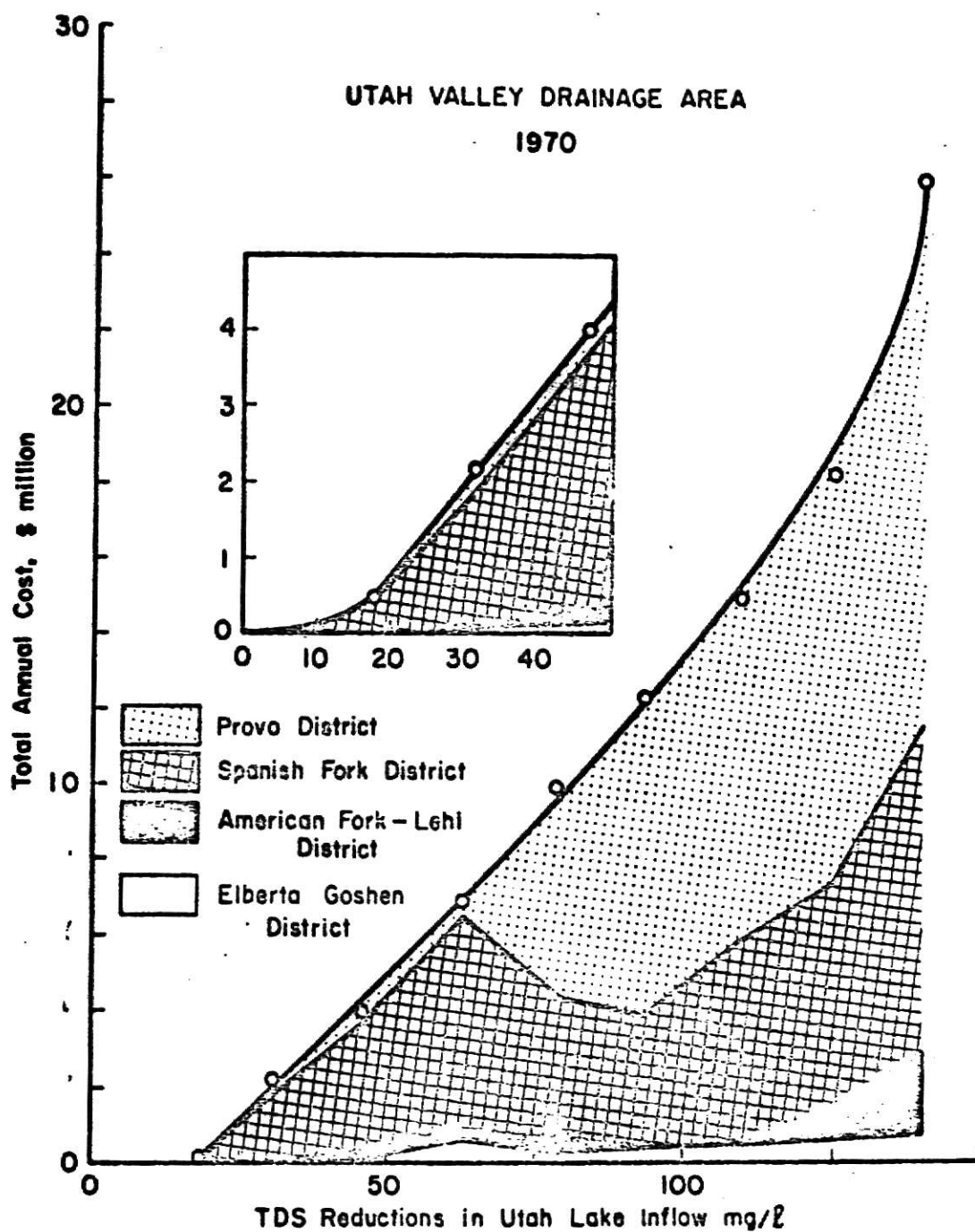


Fig. 14. Optimal water quality management policies in the Utah Valley area (Walker, et al., 1973b).

marginal cost characteristics of the urban and agricultural control measures.

The Spanish Fork district accounts for 31% of the flows in the basin, and 34% of the salts. The district is predominantly agricultural in nature. It is quite apparent, therefore, that water quality management in the district should concentrate primarily on the agricultural sector. A study of Figs. 12 and 13 substantiates this conclusion. Figure 12 is a plot of the total annual salinity control costs in the district, versus the TDS reduction standard; Fig. 13 is a plot of TDS reduction in the sector effluents versus TDS reduction in the district's effluents. Referring to Fig. 12, it is observed that the curve has two points where it changes slope. At a TDS reduction of about 75 mg/l, the noticeable break in the curve indicates a shift from control measure alternatives in the agricultural sector to a combination of structural and practice improvements. The undulation at around a TDS reduction of 130 mg/l indicates that, at this point urban desalting measures are undertaken. In fact, implementation of desalting is almost immediate for all of the urban effluent flows. Beyond this point, agricultural control measures are once again resorted to, as all of the urban effluents have already been desalted.

A comparison of the results for the Provo and the Spanish Fork districts emphasizes the need to determine optimal policies for each district. Conclusions drawn from one district are not uniformly applicable to the conditions of other districts. However, it might be tentatively stated that, for low TDS reduction in the district's effluents,

the best investment is likely to be in the agricultural sector. It is also quite obvious that if complete control is desired, both urban and agricultural salinity control measures should be undertaken in their entirety. In the middle ground, the optimal decisions are less apparent, and must be evaluated for each set of conditions.

Having optimized the individual districts policies, for a specified set of effluent standards, the next major issue is to determine an optimal policy for water quality management, under the assumption of regional operation. The development of the Regional Model has been dealt with in the preceding section. Referring to Fig. 14, it is observed that low levels of TDS reduction, up to about 70 mg/l, the bulk of the control measures are undertaken in the Spanish Fork district. It is recalled, that when the TDS reductions required are low, low marginal costs in the agricultural sector direct efforts in that direction. Beyond a TDS reduction of about 60 mg/l, it becomes more economical to employ control procedures in the Provo district, which contributes a sizeable urban effluent. It is also of interest to note that contributions to salt reduction, from the Lehi-American Fork and Elberta-Goshen districts, are insignificant. Hence, it is to be concluded that, primarily, larger areas are to be attended to first.

5.3. OPTIMAL POLICIES USING SEPARABLE PROGRAMMING

Referring to Figs. 11 through 14, it is observed that the total annual costs obtained by Walker, et al., are very much higher than the values obtained earlier in this report for the same problem considering the three additional flows, Q_y , Q_{se} , and Q_w . This does not seem

intuitively appealing. Some difference is to be expected in view of the fact that Q_y , Q_{se} , and Q_w have been ignored, as also because of the fact that the cost figures obtained by Walker, et al., are in terms of 1970 dollars. The cost figures obtained by Shojalashkari and the author are in terms of 1974 dollars.

The problem as solved by Walker, et al., was optimized using separable programming. A BOD standard of 25 mg/l was assumed, and the TDS reduction standard was artificially varied. The total annual cost figures (obtained by assuming salinity control cost for a TDS reduction of zero mg/l is zero dollars) were plotted against TDS reduction in the district's effluents (Figs. 15, 17 and 19). Figures 16 and 18 are plots of TDS reduction in the urban and agricultural sectors, against TDS reduction in the district, for the Provo and Spanish Fork districts, respectively. Suspicions about the unusually high cost figures obtained by Walker, et al., were confirmed. For example, the annual salinity control cost figure obtained by Walker, et al., for a TDS reduction of 150 mg/l in the Provo district was $\$10.40 \times 10^6$ (see Fig. 11), as against $\$1.26 \times 10^6$ for the separable programming solution (see Fig. 15). Corresponding cost figures for the Spanish Fork district for the same TDS reduction were $\$5.25 \times 10^6$ for the solution of Walker, et al. (see Fig. 12), and $\$0.35 \times 10^6$ for the separable programming solution (see Fig. 17). For the regional model, for a TDS reduction standard of 100 mg/l, Walker, et al., obtained an annual cost figure of $\$13.5 \times 10^6$ (see Fig. 13), as against $\$0.93 \times 10^6$ for the separable programming solution (see Fig. 19). Reference to Figs.

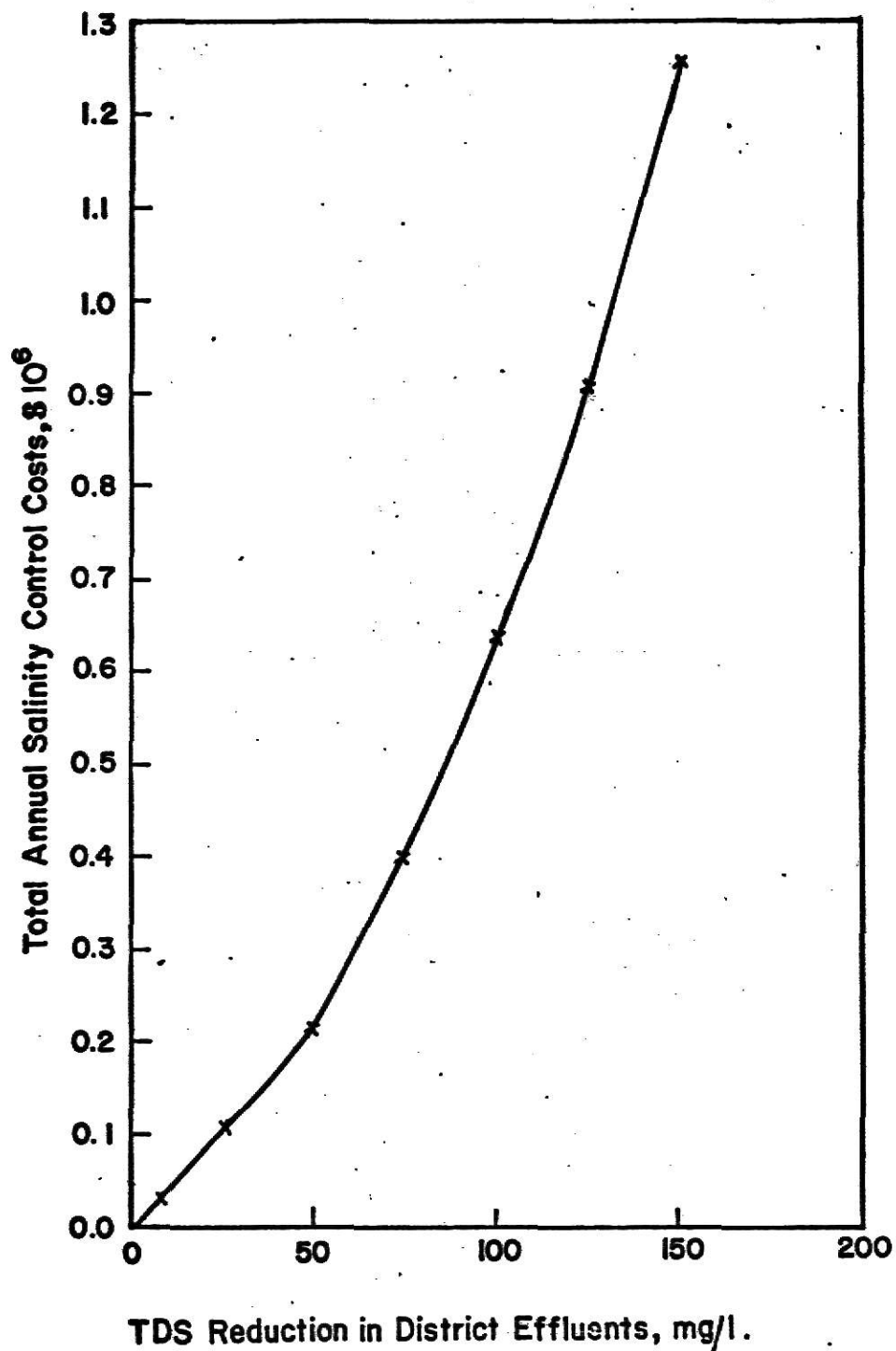


Fig. 15. Salinity control costs vs TDS reduction for the Provo District (Separable Programming Solution of Walker's version).

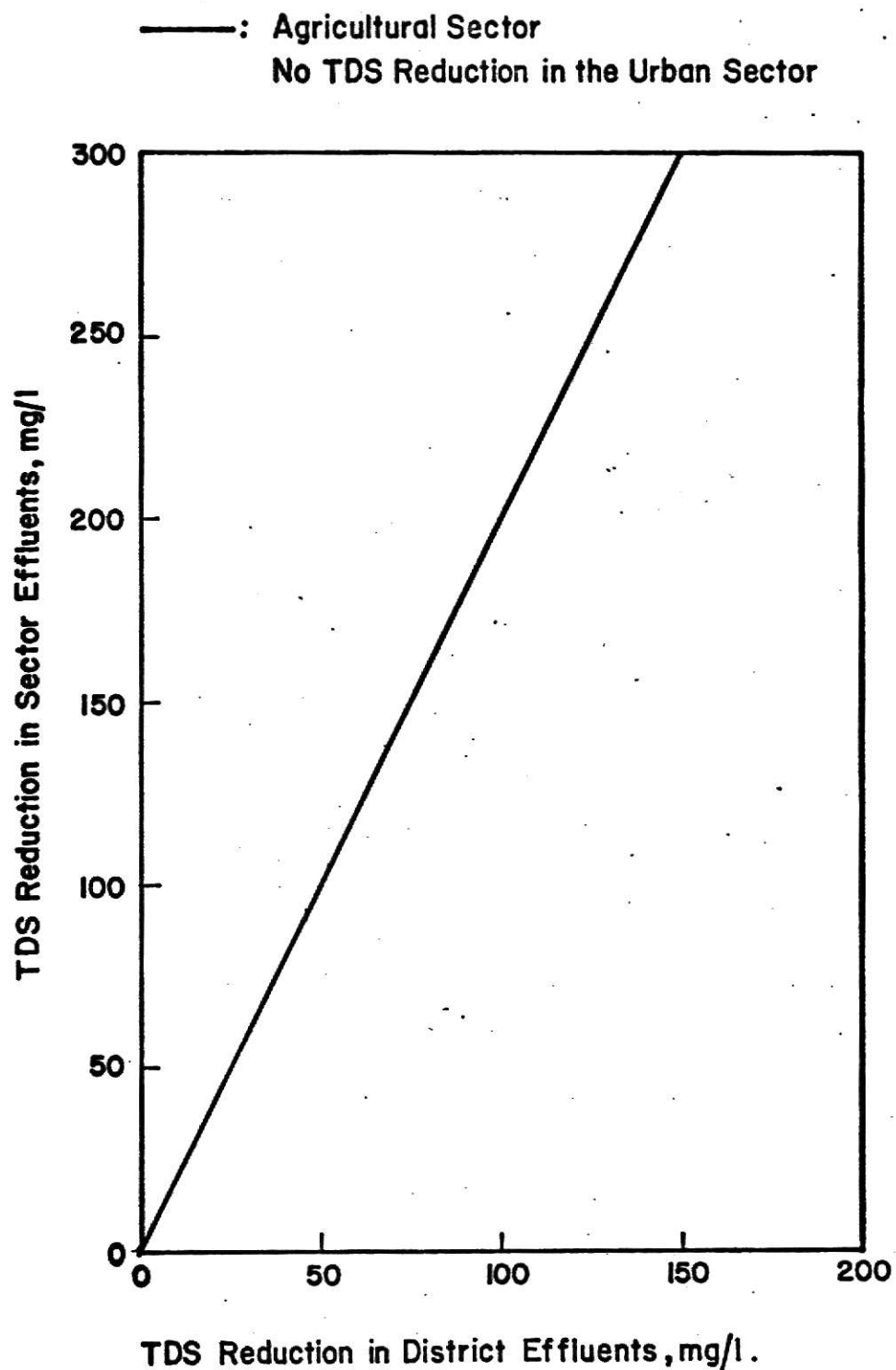


Fig. 16. TDS reduction in sectors vs TDS reduction in district effluents for the Provo District (Separable Programming Solution of Walker's version).

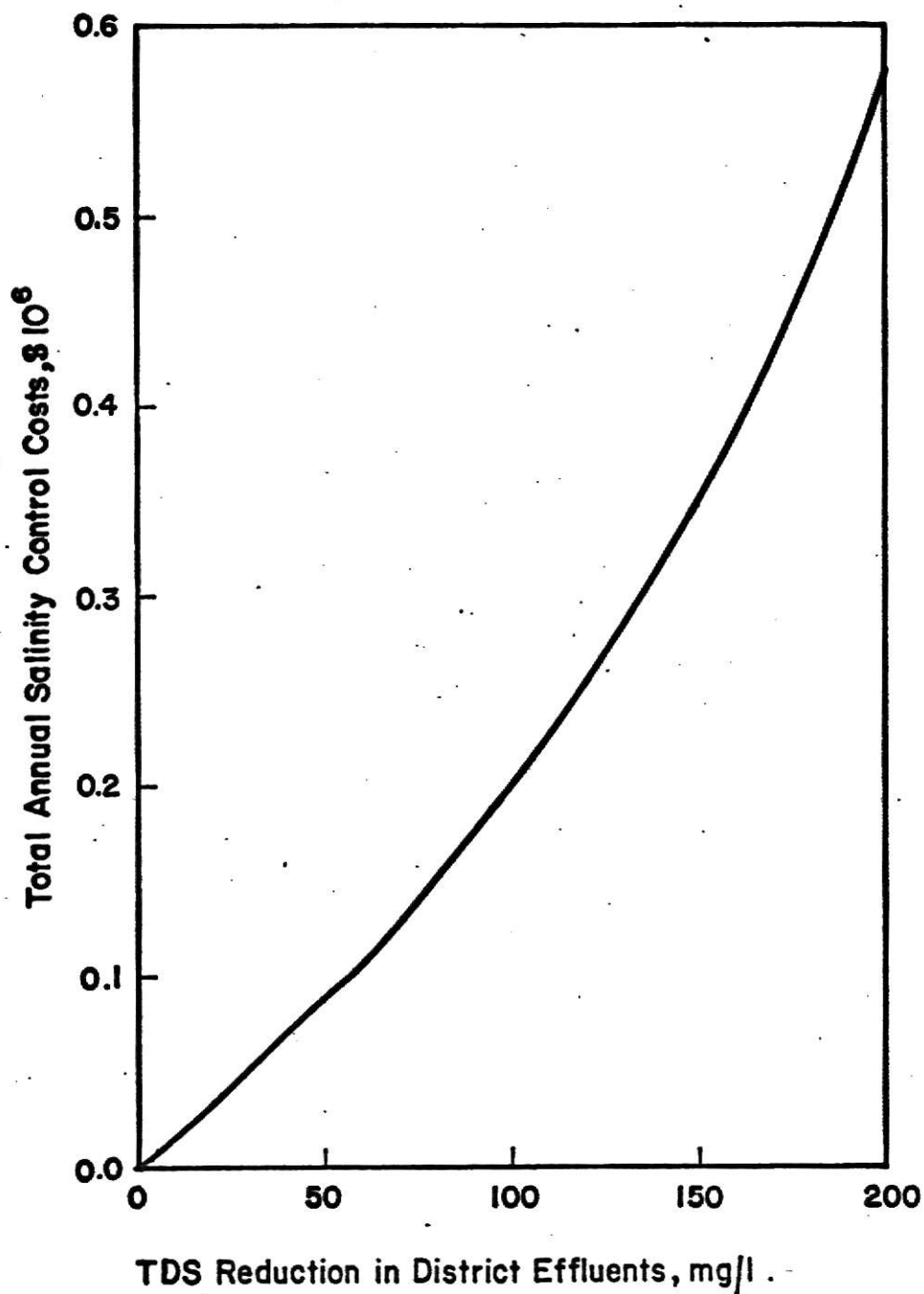


Fig. 17. Salinity control costs vs TDS reduction for the Spanish Fork District (Separable Programming Solution of Walker's version).

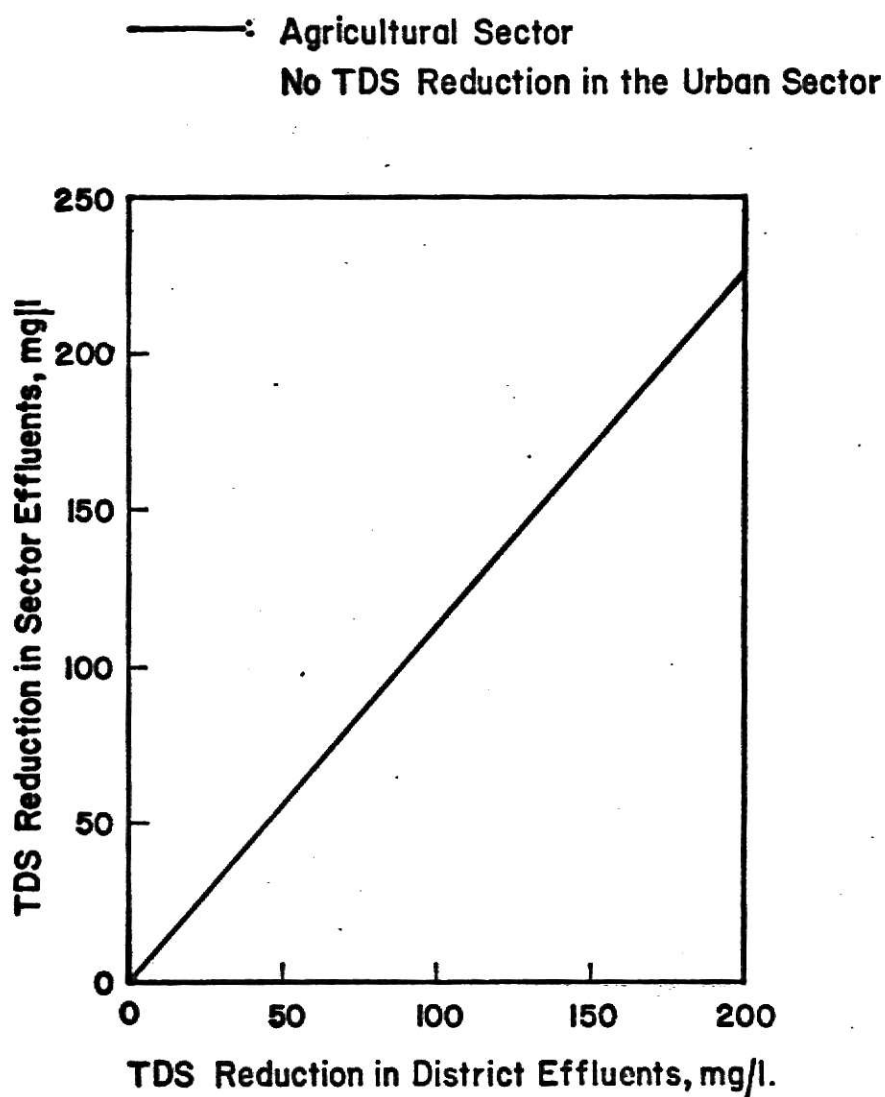


Fig. 18. TDS reduction in sectors vs TDS reduction in district effluents for the Spanish Fork District (Separable Programming Solution of Walker's version).

BOD Standard = 25 mg/l

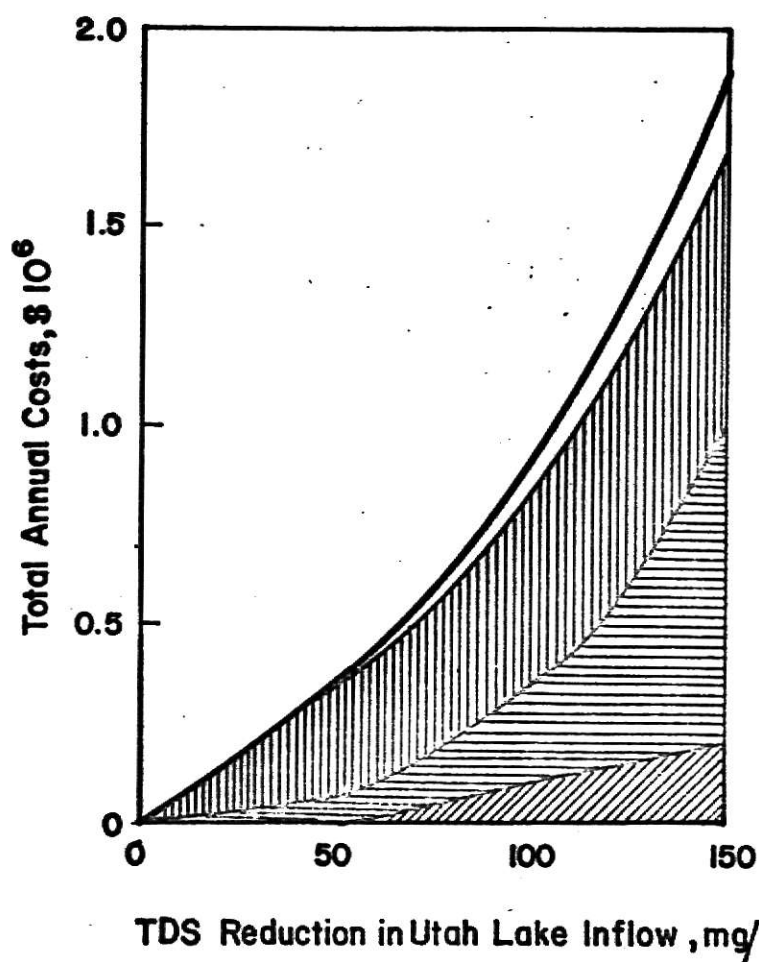
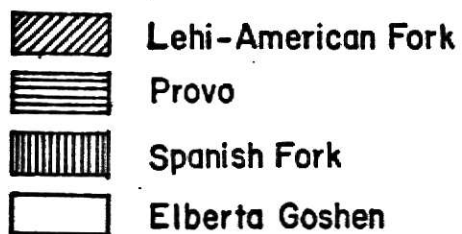


Fig. 19. Optimal water quality management policies in the Utah Valley area (Separable Programming Solution of Walker's version).

15 and 17 (separable programming solution), also reveals that for all the TDS reduction standards considered, agricultural salinity control measures are sufficient to meet the standards in both the districts. In the solution obtained by Walker, et al., in the Provo district urban salinity control measures begin at salt reduction levels as low as 25 mg/l (see Fig. 11), in the Spanish Fork district all the urban waste water is desalted for TDS reductions of 130 mg/l or higher in the district's effluents. Examining Figs. 14 and 19, it may also be inferred that there are significant differences in the ratio of cost breakdown among the four districts for the two methods of solution. As further details of the optimal policies obtained by Walker, et al., were not available, a more detailed comparison of the two solutions could not be made.

The following arguments also tend to confirm that Walker's policies are far from optimal. Referring to Table 4, it may be shown that a TDS reduction of 200 mg/l in the Spanish Fork district can be achieved by structural improvements alone (though this is not the most efficient way), at a cost of less than $\$1.3 \times 10^6$ (see Table 5). Walker's cost figure of over $\$16.0 \times 10^6$ is obviously not optimal. Similarly, in the Provo district 100% structural and practice improvements can achieve more than a TDS reduction of 150 mg/l at a cost of $\$3.9 \times 10^6$, whereas Walker's cost figure for a TDS reduction of 150 mg/l is nearly $\$11.0 \times 10^6$ (see Tables 4 and 5).

CHAPTER 6

DISCUSSION AND CONCLUDING REMARKS

This report was primarily concerned with the optimization of a water quality model, in which both urban and agricultural return flows are considered. The Utah Lake drainage area served as a source for the numerical data required. Hydrologically, the area was divided into four districts, in each of which, urban and agricultural activities exist. The Spanish Fork and the Elberta-Goshen districts are primarily agricultural in nature. The Lehi-American Fork district and the Provo district are characterized by a rapid rate of urbanization.

It was also stressed that, in areas such as the Utah Valley, where a number of polluters contribute to deterioration of water quality, water quality management strategies should aim at coordination of pollution control strategies among all the users.

Optimal policies were first determined under the assumption that each district was to meet certain BOD and TDS reduction standards. The optimal policies obtained indicated that, without exception, under present conditions, primary and secondary treatment of urban waste waters is more than adequate to meet the BOD standards. BOD reduction in the tertiary treatment unit, or BOD reduction of agricultural flows (for which the model considered in this report has no provision), was not necessary. From the point of view of salinity control, it was determined that agricultural control measures were more economical.

In the next chapter, optimal policies were determined for the situation where the salinity-BOD control activities of the districts are coordinated. The explicit purpose was to take advantage of the economies of operating on a large scale. Salt reduction standards were imposed on the aggregated return flows into the Lake, instead of on the individual districts return flows. The same could not be done with the BOD standard because existing regulations require primary and secondary treatment of all urban and industrial wastes. The total annual cost figures, under the assumption of regional operation, were \$1.824 million less per year than when the districts operated separately. Urban waste treatment costs, being the same in both cases, the savings realized are in the agricultural sector. The results obtained were compared with those of Shojalashkari (1974), using the Generalized Reduced Gradient Method.

Finally, the problem as solved by Walker, et al., (1973b), was optimized using Separable Programming. The essential difference between the problems as treated by Walker, et al., and Shojalashkari, is that the former have not considered the flows Q_y , Q_{se} and Q_w , and the associated salt concentrations. Comparison of the results obtained with those of Walker, et al., showed drastic differences in both cost figures and optimal policies.

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**APPENDIX 1. COMPUTER PROGRAM FOR GENERATING SEPARABLE PROGRAMMING
DATA IN MPS/360 FORMAT.**

```

C.....JEREL WILLIAMS.....JULY,1971.....
C
C.....THIS ROUTINE IS DESIGNED TO GENERATE THE LINEAR APPROXIMATION VALUES
C AND TO PUNCH DATA CARDS WITHIN THE SPECIFICATIONS OF MPS/360
C LINEAR AND SEPARABLE PROGRAMMING.
C
C.....A USER DESIGNED SUBROUTINE FUNC MUST ACCOMPANY THIS ROUTINE.
C
C.....THE USER MUST SUPPLY DATA FOR SIX VARIABLES. ALL DATA IS ECHU CHECKED.
C IFUN=THE NUMBER OF VARIABLES
C NROWS=THE NUMBER OF CONSTRAINTS+1
C START=THE INITIAL STARTING POINT FOR EACH VARIABLE
C INTER=THE NUMBER OF UNIFORM PARTITIONS DESIRED BETWEEN ALONG UPPER
C ALPH=THE LOWER BOUND OF THE UNIFORM REGION
C UPPER=THE UPPER BOUND OF THE UNIFORM REGION
C
C.....THE DATA CARD FORMAT IS AS FOLLOWS.
C FIRST CARD
C READ NFUN & NROWS, FORMAT 215
C THERE LOW OCCURS ONE DATA CARD FOR EACH VARIABLE CONTAINING
C START, INTER, ALPH, & UPPER, FORMAT F10.0, I10, 2F10.0
C
900 FORMAT(5X,5HEPDR/5X,30ALOW EQUALS UPPER FOR VARIABLE,15)
901 FORMAT(5X,5HEPDR/5X,31INVALID VALUE FOR NFUN (R NROWS)
902 FORMAT(5X,5HEPDR/5X,27INTER EQUALS C FOR VARIABLE,15)
1000 FORMAT(215)
2000 FORMAT(5X,1H0,14,5X,3HPDR,13,4X,F12.3,3X,3HROW,13,4X,F12.3)
3000 FORMAT(4X,1HP,14,5X,4HGRIC,13,3X,F12.1)
4000 FORMAT(4X,1HP,14,5X,3HPCN,13,4X,F12.3,3X,4HGRIC,13,3X,F12.3)
5000 FORMAT(5X,3HSET,13,4X,4HMARKER,17X,8H*SEPDIG*)
6000 FORMAT(5X,27FCHK) CHECK FOR VARIABLE,15/5X,6HSTART=F10.2/
15X,6HINTER=F10.2/5X,6HALOW =F10.2/5X,6HUPPER=F10.2/
7000 FORMAT(4X,6HENSEI,4X,4HMARKER,17X,8H*SEPDIG*/80(1H*))
8000 FORMAT(1X,14HP SPROUND P,14, 9X,3H1.0)
9000 FORMAT(5X,27FCHK) CHECK FOR NFUN & NROWS/
15X,6HNFUN =,15/5X,6HNROWS=,15/)
C
C.....UNIT NUMBERS
C
AHEAD=5
NPOINT=6
NCH=7
C
C.....NREAD=CARD INPUT,NPRINT=PRINT OUTPUT,NENCH=PUNCH OUTPUT
C
C TIMEASION PCW(53),F150,501
C READ(1PEAD,1000) NFUN,NROWS
C WRITE(1PRINT,9000) NFUN,NROWS
C IF(NFUN.LT.1.CR.NROWS.LT.1) GO TO 96
C NFK=0
C KFK=1000
C
C.....ON CALCULATIONS FOR EACH VARIABLE
C
9070 JOKK=1,NFUN
9071 IAV=1
9072 PFCAD(NPFCAD,5000) START,INTER,ALOW,UPPER
9073
9074
9075

```


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15/58/21

DATE = 76016

MAIN

FORTRAN IV G LEVEL 21

```

0026      WRITE(NPRINT,6000) KK,START,INTER,ALOW,UPPER
C.....CHECK DATA FOR ERRORS.  PROGRAM DOES NOT ALLOW INTER .GT. 50.
C

```

```

0027      IF(INTER.GT.50) INTER=50
0028      IF(INTER.EQ.0) GO TO 97
0029      IF(ALOW.EQ.UPPER) GO TO 98
C

```

```

C.....CALCULATE UNIFORM PARTITION STEP-SIZE
C

```

```

0030      AN=INTER
0031      DELTA=(UPPER-ALOW)/AN
0032      LL=KK+100
0033      WRITE(NPUNCH,4000) LL
0034      AN=NROWS+KK
C

```

```

C.....CHECK ALOW,NE,START.  CREATE ANOTHER PARTITION IF TRUE
C

```

```

0035      IJ=1
0036      IF(ALOW-START)2,3,2
0037      2 X=START
0038      AX=ALOW
0039      IJ=1
0040      INTER=INTER+1
0041      IAX=2
0042      CALL K=1,NROWS
0043      11 ROW(IK)=0.0
0044      GO TO 8
C

```

```

C.....CALCULATE PARTITION VALUES
C

```

```

0045      3 IJ=IJ+1
0046      ROW(IK)=1,NROWS
0047      10 ROW(IK)=0.0
0048      KK=IJ-IAK
0049      X=KK*DELTA+ALOW
0050      AX=X+DELTA
0051

```

```

C.....CALCULATE FUNCTION DIFFERENCES
C

```

```

0051      8 ROW(IJ)=1,NROWS
0052      CALL FUNC(J,KK,AX,F,NFUN,NROWS)
0053      ROW(IJ)=F(J,KK)
0054      CALL FUNC(J,KK,X,F,NFUN,NROWS)
0055      100 ROW(IJ)=F(J,KK)-F(J,KK)
C

```

```

C.....CREATE GRID VALUE
C

```

```

0056      GRID=X-AX
0057      II=IJ+KKK
C

```

```

C.....START PUNCH LOOP
C

```

```

0058      IP=1
0059      16 IP=IP+1
0060      IF(IP.GT.NROWS) GO TO 90
0061      IF(ROW(IP))17,200,17
0062      17 1022=IP+1

```

PAGE 0003

15/58/21

DATE = 76016

MAIN

FORTRAN IV G LEVEL 21

```

0063      GO TO 20
0064      19 IPR=IR+1
0065      20 IF(IIPR.GT.AQOKS) GO TO 89
0066      IF(RCW(IIPR)) 75,15,75
0067      89 IPR=IP+100
0068      WRITE(NPNC,H,3001) II,IRR,RCW(IRR),LL,GRID
0069      IPR=IP
0070      GO TO 700
0071      90 IPR=IP
0072      WRITE(NPNC,H,3000) II,LL,GRID
0073      GO TO 200
0074      75 IPR=IP+100
0075      IPR=IPR+100
0076      WRITE(NPNC,H,2000) II,IRR,RCW(IRR),IMM,RCW(IRR)
0077      199 IPR=IPR
0078      200 IF(II.LT.II.NN) GO TO 16
0079      IF(II.LT.II.NN) GO TO 3
0080      KK=KKK+INTER
0081      KK=KKK+INTER
0082      400 CONTINUE
0083      WRITE(NPNC,H,7000)
0084
0085      C-----CREATE RUNDS SECTION
0086      C
0087      C711=1,KKK
0088      II=I+1000
0089      1 WRITE(NPNC,H,8000) II
0090      STOP
0091      96 WRITE(NPNC,H,901)
0092      STOP
0093      97 WRITE(NPNC,H,902) KK
0094      STOP
0095      98 WRITE(NPNC,H,903) KK
0096      STOP
0097      END

```

SUBROUTINE FUNC(I,J,X,F,NFUN,NKONS)

DATA(SIN F150,50)

PA=(J-1)*NRCS+1

GO TO(1,2,3,4,5),MM

1 F(1,1)=-1*(5.23*1000*(X)**3.89+4.16*10000*(X)**0.65+

14.16*1000*(X)**0.90+1.42*10000*(X)**C.962+4.75*10000*(X)**0.72

1*5.70*1000*(X)**0.96)

RETURN

2 F(1,1)=-1*(5.34*10000*(X)**C.67+22.76*10000*(X)**0.75)

RETURN

3 F(1,1)=-1*764840*34.42*(3.68135-ALOG(39.70-(X)))/100.00

RETURN

4 F(1,4)=-1*2544270*37.60*(4.11087-ALOG(56.50-(X)))/100.00

RETURN

5 F(1,5)=-1*(3.31*10000*(X)**0.71+2.12*10000*(X)**0.83+

16.67*10000*(X)**0.80+4.29*10000*(X)**0.893)

RETURN

END

0001

0002

0003

0004

0005

0006

0007

0008

0009

0010

0011

0012

0013

0014

0015

APPENDIX 2. COMPUTER OUTPUT OF A TYPICAL SEPARABLE PROGRAMMING PROBLEM.

The optimization problem solved is for the Lehi-American Fork district.

BOD standard = 25 mg/l

TDS reduction standard = 150 mg/l

CONTROL PROGRAM COMPILER - MPS/360 V2-M11

```

0301 PROGRAM
0302 INITIALZ
0303 MOVE(XDATA,'DATA-SET')
0304 MOVE(XOBJNAME,'EXAMPLE')
0305 CONVERT
0306 RCDCCT
0307 MOVE(XOBJ,'RCDCCT')
0308 MOVE(XHHS,'LIMITS')
0309 TITLE('PRIMAL-LOCAL OPTIMUM')
0310 XSETL=-1 $* ROUNDED VARIABLES AT LOWER LIMIT
0311 SETUP('RCUND','SEPRCUND','MAX')
0312 PRIMAL
0313 SOLUTION $* LOCAL OPTIMUM
0314 TITLE('PRIMAL-GLCHAL OPTIMUM')
0315 XSETL=+1 $* ROUNDED VARIABLES AT UPPER LIMIT
0316 SETUP('RCUND','SEPRCUND','MAX')
0317 PRIMAL
0318 SOLUTION $* GLOBAL OPTIMUM
0319 TITLE('DUAL-GLCHAL OPTIMUM')
0320 SETUP('RCUND','SEPRCUND','MAX')
0321 DUAL
0322 SOLUTION $* GLOBAL OPTIMUM DUAL
0323 EXIT
0324 PEND
0325

```

EXECUTOR. MPS/360 V2-M11

CONCEPT DATA-SET TO EXAMPLE

TIME = 0.01

1- FCHS SECTION.

0 MINOR ERROR(S) - 0 MAJOR ERROR(S).

2- COLUMNS SECTION.

0 MINOR ERROR(S) - 0 MAJOR ERROR(S).

3- RMS'S SECTION.

LIMITS

0 MINOR ERROR(S) - 0 MAJOR ERROR(S).

5- RCUES SECTION.

RCUES
SEPARING

0 MINOR ERROR(S) - 0 MAJOR ERROR(S).

PROBLEM STATISTICS - 11 RCUES, 68 VARIABLES, 132 ELEMENTS, DENSITY = 17.64

THESE STATISTICS INCLUDE ONE SLACK VARIABLE FOR EACH RCUE.

0 MINOR ERRORS, 0 MAJOR ERRORS.

[illegible]

EM-CATA

PRIMAL-LOCAL OPTIMUM

SOLUTION (OPTIMAL)

TIME = 0.24 MINS. ITERATION NUMBER = 61

...NAME...	...ACTIVITY...	DEFINED AS
FUNCTIONAL	1903809.4473R-	ROW101
RESTRAINTS		LIMITS
PCOUNDS....		SEPCOUND

PRIMAL-LOCAL OPTIMUM

SECTION 1 - ROWS

NUMBER	...RC...	AT	...ACTIVITY...	SLACK ACTIVITY	..LOWER LIMIT..	..UPPER LIMIT..	..DUAL ACTIVITY
1	00A101	RS	1904800.44778-	1904800.44738	NCNE	NCNE	1.00000
2	00A102	FC	6.42000	.	6.42000	6.42000	92255.16044
3	00A103	FC	6.42000	.	6.42000	6.42000	66921.12534
4	00A104	RS	173.24000	552.96000	NCNE	726.30000	.
5	00A105	UL	11680.70000-	.	NONE	11680.70000-	167.30291-
6	00A106	FC	92255.16044-
7	00A107	RS
8	00A108	FC	60229.01241-
9	00A109	FC	37371.74107-
10	00A110	FC	37371.74107-
11	00A111	FC	92255.16044-

SECTION 2 - COLUMNS

PIOTAL-LOCAL OPTIMUM

NUMBER	COLUMN	AT	ACTIVITY...	INPUT COST..	LOWER LIMIT.	UPPER LIMIT.	REDUCED COST.
12	A1	LL				NCNE	
13	A2	RS	6.42000			NCNE	
14	A3	RS				NCNE	
15	A4	RS				NCNE	
16	A5	RS	33.75000			NCNE	
17	A6	RS	30.03744			NCNE	
18	A7	RS	6.42000			NCNE	
19	P1021	LL		121631.18800-		1.00000	121631.18800-
20	P1022	LL		79456.09000-		1.00000	79456.09000-
21	P1023	LL		65289.62500-		1.00000	65289.62500-
22	P1024	LL		63452.56300-		1.00000	63452.56300-
23	P1025	LL		59552.31300-		1.00000	59552.31300-
24	P1026	LL		56418.68800-		1.00000	56418.68800-
25	P1027	LL		54308.19600-		1.00000	54308.19600-
26	P1028	LL		52420.25000-		1.00000	52420.25000-
27	P1029	LL		50433.56300-		1.00000	50433.56300-
28	P1030	LL		48472.25000-		1.00000	48472.25000-
29	P1031	LL		200052.14000-		1.00000	161365.15178-
30	P1032	LL		163376.68800-		1.00000	101739.66178-
31	P1033	LL		124431.51300-		1.00000	85736.47378-
32	P1034	LL		11501.50300-		1.00000	76414.47178-
33	P1035	LL		108637.53000-		1.00000	69567.47378-
34	P1036	LL		103755.25000-		1.00000	65048.22378-
35	P1037	LL		55817.50000-		1.00000	61210.47378-
36	P1038	LL		96676.15000-		1.00000	59009.16178-
37	P1039	LL		53664.75000-		1.00000	55297.72778-
38	P1040	LL		91635.00000-		1.00000	52557.07178-
39	P1041	UL	1.00000	24071.26700-		1.00000	116072.16202
40	P1042	UL	1.00000	26726.11000-		1.00000	114917.31672
41	P1043	UL	1.00000	30034.41000-		1.00000	110105.21672
42	P1044	UL	1.00000	34291.125000-		1.00000	105952.90402
43	P1045	UL	1.00000	35564.68000-		1.00000	100195.34102
44	P1046	UL	1.00000	47844.47500-		1.00000	92205.15402
45	P1047	UL	1.00000	55664.70000-		1.00000	87480.32402
46	P1048	UL	1.00000	79255.67500-		1.00000	60847.40402
47	P1049	UL	1.00000	118509.07500-		1.00000	21574.05102
48	P1050	UL	1.00000	241376.50000-		1.00000	101232.47358-
49	P1051	UL	1.00000	55852.43800-		1.00000	100426.41200
50	P1052	UL	1.00000	111502.47500-		1.00000	97774.17500
51	P1053	UL	1.00000	126234.25000-		1.00000	83347.53300
52	P1054	UL	1.00000	145641.56300-		1.00000	63070.18700
53	P1055	UL	1.00000	171616.10000-		1.00000	37665.56200
54	P1056	AS	36383	209201.75000-		1.00000	58989.91300-
55	P1057	LL		270271.56300-		1.00000	164877.25000-
56	P1058	LL		374159.00000-		1.00000	413774.25000-
57	P1059	LL		623061.00000-		1.00000	214375.25000-
58	P1060	LL		239257.00000-		1.00000	51372.37500-
59	P1061	UL	1.00000	11000.18800-		1.00000	23456.18700-
60	P1062	UL	1.00000	83084.00000-		1.00000	

ORIGINAL-LOCAL OPTIMUM

NUMBER	CELLNO.	AT	...ACTIVITY...	..INPUT COST..	..LOWER LIMIT.	..UPPER LIMIT.	..REDUCED COST.
61	P1043	UL	1-CC000	75430.43800-	.	1.00000	16202.62500-
62	P1044	UL	1-00300	70896.43800-	.	1.00000	11468.62500-
63	P1045	UL	1-CC000	67725.18800-	.	1.00000	6457.37500-
64	P1046	UL	1-CC000	65312.43800-	.	1.00000	6084.75000-
65	P1047	UL	1-00300	63379.93800-	.	1.00000	4152.12500-
66	P1048	UL	1-CC000	61775.93800-	.	1.00000	2548.12500-
67	P1049	UL	1-00300	60411.00300-	.	1.00000	1183.18700-
68	P1050	BS	1-00300	59227.81300-	.	1.00000	.

PRIMAL-GLOBAL OPTIMUM

SOLUTION (OPTIMAL)

TIME = 0.36 MINS. ITERATION NUMBER = 45

...NAME...	...ACTIVITY...	DEFINED AS
SUPPLEMENTAL	150809.44738-	RM1C1
RESTRAINTS		LIMITS
ACLVNS...		SEPCUND

PRIMAL-GLOBAL OPTIMUM

SECTION 1 - ACBS

NUMBER	...SYM...	AT	...ACTIVITY...	SLACK ACTIVITY	..LOWER LIMIT.	..UPPER LIMIT.	..DUAL ACTIVITY
1	PM121	BS	190809.44738-	190809.44738	NONE	NONE	1.07000
2	PM122	EQ	6.42000	.	6.42000	6.42000	92255.16044
3	PM123	EQ	6.42000	.	6.42000	6.42000	66921.12534
4	PM124	FS	173.74000	552.96000	NONE	726.73000	.
5	PM125	UL	11640.70000-	.	NONE	11680.70000-	167.30281-
6	PM126	EQ	92255.16044-
7	GM127	FS	60229.01281-
8	GM128	EQ	37371.74107-
9	GM129	EQ	37371.74107-
10	GM130	EQ	92255.16044-
11	GM131	EQ

SECTION 2 - COLUMNS
PARTIAL-GLOBAL OPTIMUM

NUMBER	COLUMN	AT	ACTIVITY...	INPUT COST...	LOWER LIMIT	UPPER LIMIT	REDUCED COST
12	X1	LL				NNNN	
13	X2	NS	6.42000			NNNN	
14	X3	NS				NNNN	
15	X4	NS				NNNN	
16	X5	NS	33.15000			NNNN	
17	X6	NS	30.03744			NNNN	
18	X7	NS	6.42000			NNNN	
19	P1001	LL		121431.18000		1.00000	121431.18000
20	P1002	LL		79405.92000		1.00000	79405.92000
21	P1003	LL		69289.62500		1.00000	69289.62500
22	P1004	LL		63462.56300		1.00000	63462.56300
23	P1005	LL		59552.31300		1.00000	59552.31300
24	P1006	LL		56618.68800		1.00000	56618.68800
25	P1007	LL		54303.75000		1.00000	54303.75000
26	P1008	LL		52420.25000		1.00000	52420.25000
27	P1009	LL		50833.56300		1.00000	50833.56300
28	P1010	LL		49472.25000		1.00000	49472.25000
29	P1011	LL		200052.18800		1.00000	161385.16174
30	P1012	LL		143376.68800		1.00000	131700.66174
31	P1013	LL		124403.50000		1.00000	85736.47378
32	P1014	LL		115081.50000		1.00000	76414.47378
33	P1015	LL		109677.50000		1.00000	69500.47378
34	P1016	LL		103755.75000		1.00000	65349.22378
35	P1017	LL		55877.50000		1.00000	61210.47378
36	P1018	LL		96674.18800		1.00000	54009.16174
37	P1019	LL		93764.75000		1.00000	55297.72378
38	P1020	LL		91625.00000		1.00000	52457.47378
39	P1021	LL	1.00000	74371.36700		1.00000	116072.16700
40	P1022	UL	1.00000	26726.11100		1.00000	113417.01602
41	P1023	UL	1.00000	30334.81100		1.00000	110105.21602
42	P1024	UL	1.00000	34291.12500		1.00000	105852.90602
43	P1025	UL	1.00000	39544.68800		1.00000	100195.34102
44	P1026	UL	1.00000	47444.87500		1.00000	87255.15402
45	P1027	UL	1.00000	53664.00000		1.00000	80680.72402
46	P1028	UL	1.00000	70295.62500		1.00000	60647.40402
47	P1029	UL	1.00000	114569.91600		1.00000	21574.09102
48	P1030	UL		241376.50000		1.00000	131232.47098
49	P1031	UL		29552.61300		1.00000	134429.31200
50	P1032	UL	1.00000	111537.87500		1.00000	97779.37500
51	P1033	UL	1.00000	126234.25000		1.00000	83047.50000
52	P1034	UL	1.00000	143461.56300		1.00000	63420.13700
53	P1035	UL	1.00000	171616.18800		1.00000	37665.56200
54	P1036	NS	36183	209441.75000		1.00000	
55	P1037	LL		26221.56300		1.00000	59944.41300
56	P1038	LL		374159.00000		1.00000	164877.25000
57	P1039	LL		623051.00000		1.00000	413775.25000
58	P1040	LL		2392957.00000		1.00000	2183675.25000
59	P1041	UL	1.00000	110303.18800		1.00000	51072.37500
60	P1042	UL	1.00000	33044.00000		1.00000	23456.12700

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PRIMAL-GLOBAL CPT IMP

NUMBER	COLUMN	AT	ACTIVITY...	INPUT COST...	LOWER LIMIT.	UPPER LIMIT.	REDUCED COST.
61	P1343	UL	1.CC000	75433.43800-	.	1.00000	16202.62900-
62	P1044	UL	1.00000	70896.47800-	.	1.00000	11664.62500-
63	P1345	UL	1.CC000	67725.18000-	.	1.00000	3487.17500-
64	P1046	UL	1.00000	65312.52300-	.	1.00000	6084.75000-
65	P1047	UL	1.CC000	63379.93100-	.	1.00000	4152.12500-
66	P1048	UL	1.CC000	61775.92800-	.	1.00000	2548.12500-
67	P1347	UL	1.CC000	67411.03300-	.	1.00000	1183.18700-
68	P1053	PS	1.CC000	59227.81300-	.	1.00000	.

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**APPLICATION OF SEPARABLE PROGRAMMING
TO REGIONAL WATER QUALITY MANAGEMENT**

by

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AN ABSTRACT OF A MASTER'S REPORT

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Walker, et al., have developed a mathematical model of water quality management that optimizes allocation of water pollution control strategies in an area where both urban and agricultural activities exist side by side. The model was applied to the Utah Lake drainage area which was subdivided into four districts, each of which discharges urban and agricultural return flows into the Lake. Optimal policies that minimize the cost of maintaining specific BOD and salinity standards were first determined under the assumption of independent operation among the districts. Next it was assumed that the salinity-BOD control efforts of the districts could be regionally coordinated to take advantage of the economies of operating on a larger scale. As before BOD standards were imposed on the effluents from each district, but TDS reduction standards were imposed on the aggregated return flows. The method of optimization used by Walker, et al., was the 'Jacobian differential algorithm'. Shojalashkari used the generalized reduced gradient method to optimize a slightly different version of the same system.

This study was undertaken because of serious discrepancies among the solutions to the two versions mentioned above. Separable programming which is a comparatively simpler nonlinear programming technique, was used as the method of optimization in this study. A simple mathematical manipulation was used to convert the objective function to a form amenable to separable programming. The separable programming option in MPS/360 was used to resolve the problem as solved by Walker, et al., and Shojalashkari. The optimal policies

obtained in this study agree well with those determined by Shojalashkari. However, there was total disagreement with the optimal policies proposed by Walker, et al.